



Article title: Light pollution: A landscape-scale issue requiring cross-realm consideration

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06th May 2022

Professor Craig Styan
Editor
UCL Open: Environment



Dear Professor Styan,

We would like to thank you and the reviewers for the critical evaluation and constructive comments on our manuscript entitled: "Light pollution: A landscape-scale issue requiring cross-realm consideration".

We found the reviewers' comments very helpful and have modified the manuscript following their suggestions. We believe that the modifications have greatly improved the quality of the manuscript.

We have included two revised versions of our manuscript for publication in UCL Open: Environment: a version in which the changes have been tracked and a clean version of the revised manuscript.

We have outlined our detailed responses to reviewers in a different file (also attached).

We hope that our changes will now make the paper acceptable for publication.

Yours Sincerely,

A handwritten signature in black ink, appearing to read 'Mariana Mayer Pinto'. The signature is written in a cursive style and is positioned above a horizontal line.

Dr Mariana Mayer Pinto
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1 **Light pollution: A landscape-scale issue requiring cross-realm consideration**

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28 **ABSTRACT**

29

30 Terrestrial, marine, and freshwater realms are inherently linked through ecological,
31 biogeochemical and/or physical processes. An understanding of these connections is critical
32 to optimise management strategies and ensure the ongoing resilience of ecosystems. Artificial
33 light at night (ALAN) is a global stressor that can profoundly affect a wide range of
34 organisms and habitats and impact multiple realms. Despite this, current management
35 practices for light pollution rarely consider connectivity between realms. Here we discuss the
36 ways in which ALAN can have cross-realm impacts and provide case studies for each
37 example discussed. We identified three main ways in which ALAN can affect two or more
38 realms: 1) impacts on species that have life cycles and/or stages in two or more realms, such
39 as diadromous fish that cross realms during ontogenetic migrations and many terrestrial
40 insects that have juvenile phases of the lifecycle in aquatic realms; 2) impacts on species
41 interactions that occur across realm boundaries, and 3) impacts on transition zones or
42 ecosystems such as mangroves and estuaries. We then propose a framework for cross-realm
43 management of light pollution and discuss current challenges and potential solutions to
44 increase the uptake of a cross-realm approach for ALAN management. We argue that the
45 strengthening and formalisation of professional networks that involve academics, lighting
46 practitioners, environmental managers and regulators that work in multiple realms is essential
47 to provide an integrated approach to light pollution. Networks that have a strong multi-realm
48 and multi-disciplinary focus are important as they enable a holistic understanding of issues
49 related to ALAN.

50

51 **KEY-WORDS:** ALAN, artificial light at night, light pollution, multi-disciplinary, adaptive
52 management, ecological connectivity.

53 INTRODUCTION

54 Artificial light at night (ALAN) is a widespread anthropogenic pollutant that is
55 rapidly increasing in intensity and global distribution. Current estimates suggest more than
56 80% of the human population, and nearly a quarter of the global land area, are exposed to
57 light-polluted skies (Falchi et al. 2016). Consequently, ALAN affects most ecosystems
58 globally, with the potential for profound impacts. At its core, ALAN alters natural light-dark
59 cycles, disrupting a key driver of biological, ecological and evolutionary processes (Gaston et
60 al. 2014, Hopkins et al. 2018). Emergent research has linked the presence of ALAN to altered
61 physiology of plants (Bennie et al. 2016) and animals (Dominoni et al. 2013); shifts in
62 activity patterns, behaviours, reproduction and survival of animals (Robert et al. 2015,
63 Sanders et al. 2020); disruption of trophic and non-trophic species interactions (Bennie et al.
64 2015, Gaston et al. 2017); and, significant changes to the structure of ecological communities
65 (Davies et al. 2015, Hölker et al. 2015). The importance and severity of potential effects of
66 this stressor are recognised across multiple taxa, habitats and ecosystems (Sanders et al.
67 2020) and there is an increased desire to devise management strategies to minimise
68 ecological impacts of ALAN.

69 A major challenge with mitigating the impacts of ALAN is that, while it is a global
70 environmental pollutant (Falchi et al. 2016) that damages ecological systems (Sanders et al.
71 2020), it is also central to the functioning of modern human society (Edensor 2017).
72 However, beyond natural systems, ALAN can pose public health risks (Pauley 2004) and is
73 energetically and economically costly (Gallaway et al. 2010). Strategies to address the
74 ecological challenges posed by ALAN therefore need to be interdisciplinary, involving
75 researchers (e.g. ecologists, physiologists, social scientists, physicists), managers or
76 regulators (e.g. local councils and government agencies), and practitioners (e.g. urban
77 planners, developers, health specialists, and lighting professionals). While interdisciplinary

78 frameworks have been developed to foster collaboration among researchers, managers and
79 practitioners to better manage urban lighting (e.g. Pérez Vega et al. 2021), they are largely
80 applied within an individual realm, e.g. marine, freshwater or terrestrial. We use the term
81 ‘realm’ as defined by Bugnot et al. (2019), to encompass a group of ecosystems that share
82 common physical and ecological attributes (e.g. the marine realm includes all ecosystems
83 present below the high tide mark while the terrestrial realm includes both air and land).
84 Although realms are often considered as separate entities, they are intrinsically linked
85 through ecological, biogeochemical and/or physical processes. Where these linkages are
86 compromised, ecosystem functioning and services are affected and ecological systems may
87 become less biodiverse and/or resilient to change (Beger et al. 2010, Field and Parrott 2017).
88 Nevertheless, current management practices for light pollution do not consider connectivity
89 between realms. The lack of a multiple-realm integrated approach means outcomes of
90 practices are likely limited, at best, to small-scale, localised and/or temporary benefits
91 (Threlfall et al. 2021).

92 In this paper, we review examples where ALAN affects two or more realms, directly
93 and/or indirectly, and provide case studies for each example discussed. We identify three
94 main ways in which ALAN can have cross-realm effects: through impacts on 1) species that
95 have life cycles and/or stages in two or more realms, such as diadromous fish that cross
96 realms during ontogenetic migrations and many terrestrial insects that have juvenile phases of
97 the lifecycle in aquatic realms; 2) species interactions that occur across realm boundaries; 3)
98 transition zones or ecosystems such as mangroves and estuaries. We discuss the
99 consequences of taking a single-realm approach to light pollution management and present a
100 framework to help bridge this gap, incorporating both theoretical and empirical
101 considerations. We also discuss existing challenges and hurdles to studying and managing
102 light pollution. Given ALAN is projected to increase in all three realms in response to

103 continuing human population growth (Kyba et al. 2017), cross-realm management will be
104 critical for ensuring the ongoing resilience of ecosystems (Threlfall et al. 2021).

105

106 **Impacts of ALAN on two or more realms**

107 *Potential paths for cross-realm impacts*

108 Mitigating the impacts of ALAN and prioritising appropriate conservation actions
109 requires consideration of the fundamental interactions among realms (e.g. terrestrial, marine
110 and freshwater) (Beger et al. 2010). Shifts in ecological connectivity through the disruption
111 of daily, seasonal or other cyclic movement of organisms or resources can have multi-realm
112 consequences. For example, variation at the level of individual or population can affect food-
113 webs directly but also influence functions such as pollination and nutrient cycling. These
114 shifts, can in turn, have cross-realm implications due to trophic cascades and linked changes
115 in ecosystem functions. This is particularly true if the organisms involved typically function
116 across realm boundaries. Similarly, individual-level shifts can have cross-realm ecological
117 consequences if the species in question has life histories or migratory patterns that traverse
118 multiple realms, such as the two case studies we discuss below, salmon (freshwater juveniles,
119 marine adults) and secondarily aquatic insects (aquatic juveniles, terrestrial adults).

120 Throughout the paper, and in each case study presented, we outline known, measured impacts
121 of ALAN, incorporate additional existing knowledge of species and/or habitats, and discuss
122 how these may influence multiple realms.

123 Demonstrated impacts of ALAN include changes in the phenology, growth form and
124 resource allocation of plants (Bennie et al. 2016), as well as the behaviour, physiology,
125 distribution and survival of animals (Brüning et al. 2011, Perkin et al. 2014, Bolton et al.
126 2017, Fobert et al. 2019, Willmott et al. 2019, Aulsebrook et al. 2020). There are multiple
127 mechanisms underpinning these observed changes which may directly or indirectly affect

128 other realms. For example, changes in the flux of inorganic and organic material (such as
129 oxygen and nutrient fluxes), can directly impact land, sea and freshwater habitats (Hölker et
130 al. 2015, Grubisic et al. 2017), while indirect effects can be driven by bottom-up or top-down
131 processes. For example, decreased diversity and abundance of aquatic insects due to ALAN
132 is expected to affect terrestrial consumers that rely on aquatic prey, such as spiders, birds and
133 bats (Baxter et al. 2005, Zapata et al. 2019). Alternatively, changes may be driven by top-
134 down processes, arising from e.g. shifts in the survival or behaviour of herbivores and/or
135 predators. The consequences of such changes are varied and magnitude-dependent, but they
136 can result in loss of biodiversity (Bowyer et al. 2005).

137 Transitional zones, such as estuaries and coastal wetlands, including the organisms
138 that inhabit them, tend to be disproportionately affected by ALAN, because urban settlements,
139 where ALAN is prevalent, are often developed near waterways (Kummu et al. 2011).
140 Moreover, as these ecosystems are at the intersection of freshwater, marine, and terrestrial
141 realms, any ALAN effects are likely to have cross-realms consequences.

142 Rapid changes in the environment, including those linked to ALAN, can alter the
143 environmental cues many animals use to select optimal habitats resulting in them selecting
144 sites that reduce their fitness (Hale and Swearer 2016, Swearer et al. 2021). These ‘ecological
145 traps’ can promote disruptions or alterations in the movement patterns of organisms, resulting
146 in increased risk of mortality and/or shifts in trophic interactions (Schlaepfer et al. 2002).
147 Ecological traps may not inherently have cross-realm impacts, however ALAN can disrupt
148 species interactions or individual movements to create an ecological trap across more than
149 one realm (see Box 1).

150 Given the above, we have identified three broad pathways ALAN can have cross-
151 realm impacts: 1) for species that move across realms, through life cycles and/or stages or
152 migratory patterns that occur in two or more realms, such as diadromous fish and many

153 insects, as well as marine reptiles (e.g. turtles), mammals (e.g. seals) and birds (e.g. penguins
154 and albatross) that are tied to land for breeding and/or resting; 2) where species interactions,
155 such as predator-prey interactions, occur across realm boundaries; and 3) at transition zones
156 or ecosystems such as coastal wetlands and estuaries, where multiple realms are inherently
157 linked. These cross-realm linkages can be further affected if ALAN acts as an ecological trap
158 (see Box 1). Below, we provide case studies of each way in which ALAN-related impacts can
159 have cross-realm consequences.

160

161 ***1) Impacts on species with life cycles/stages across two or more realms***

162 The life cycles of many organisms occur in two or more realms. Examples include
163 animals whose juveniles are aquatic while adults are predominantly marine or terrestrial, or
164 marine animals that breed on land or in freshwater systems. Impacts of ALAN on any one
165 stage are, therefore, predicted to have carry-over effects on subsequent life-stages,
166 consequently impacting different realms. We use two case studies to illustrate this, one on
167 salmon (Salmonidae) and the other gives a broader overview of secondarily aquatic insects,
168 such as dragonflies and mayflies.

169

170 ***Case study 1 – Salmon, a vector of energy and nutrients across realms***

171 ***Demonstrated ALAN impacts***

172 Salmon, including the Atlantic (*Salmo salar*) and Pacific Salmon (*Oncorhynchus* spp.),
173 are anadromous fish - they spend their juvenile phase (e.g. alevins, fry, and parr) in rivers,
174 before migrating to the ocean as smolts (1-3 yr old juveniles that are physiologically adapted
175 for sea water) to feed, grow, and mature. Adults then return to freshwater systems for
176 spawning (Figure 1). ALAN has demonstrable impacts on several life-stages of salmon
177 species including fry (Riley et al. 2013, Riley et al. 2015) and smolts (Riley et al. 2012). For

178 example, emergence of juvenile Atlantic salmon in streams is usually mediated by
179 environmental cues, such as presence of predators (Jones et al. 2003, Falcón et al. 2020). Fry
180 are highly vulnerable to predation, and synchronous emergence can increase their chance of
181 survival (Brannas 1995). However, in freshwater river systems, ALAN is linked to
182 asynchronous nocturnal emergence, disrupted dispersal and decreased weight of fry (Riley et
183 al. 2013). Experimental field evidence also demonstrated that smolt populations exposed to
184 ALAN from streetlights along their native streams altered their migratory behaviour towards
185 the sea, with potential consequences for their fitness and/or predation risk (Riley et al. 2012).
186 In the marine realm, ALAN associated with aquaculture practices, alters the vertical
187 movement of smolt, resulting in potential trade-offs between preferred light and temperature
188 levels, feeding, and risk perception (Oppedal et al. 2011). For example, surface mounted
189 lights used in commercial farming induced movement of the smolt towards the surface,
190 resulting in higher schooling densities and shallower nocturnal swimming depths compared
191 to the day. This results in suboptimal environmental conditions and crowding of fish (Juell
192 and Fosseidengen 2004), with likely consequences to their growth and survival rates.

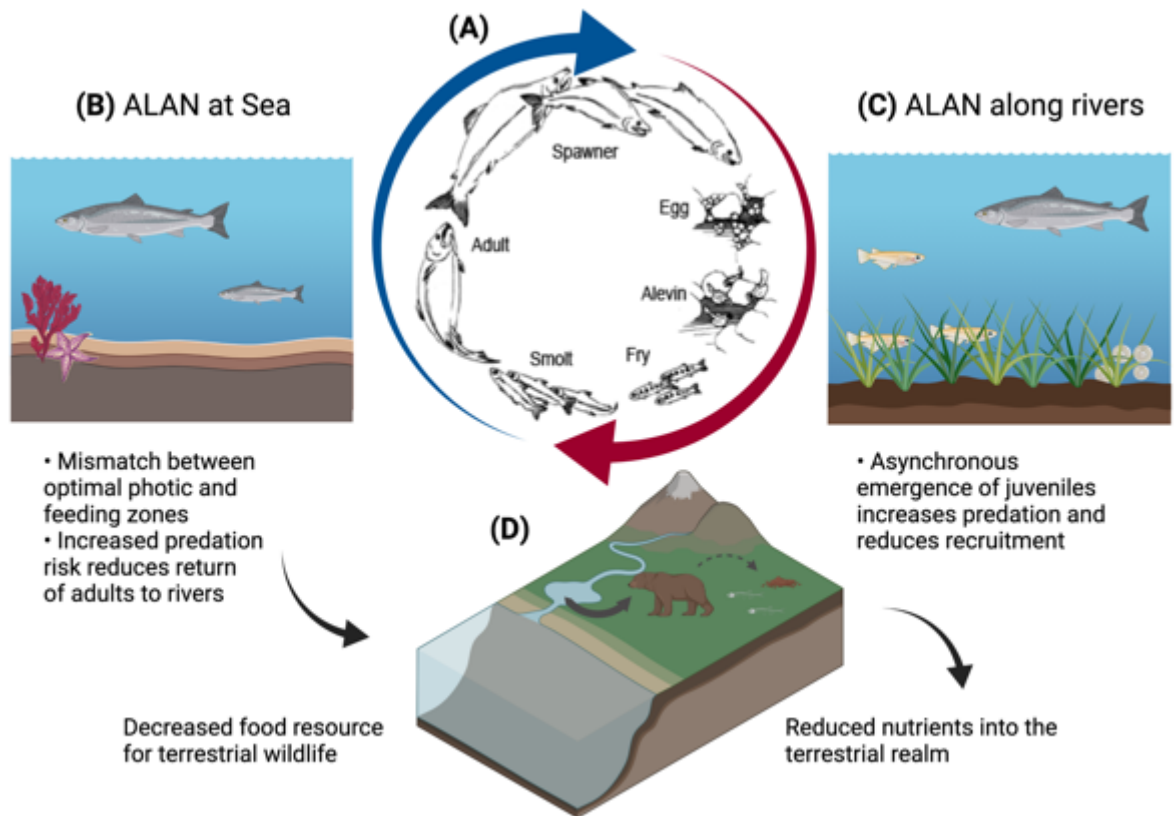
193 Potential cross-realms effects and knowledge gaps

194 Salmon are important vectors in transporting energy and nutrients between the ocean,
195 freshwater and terrestrial environments (Gende et al. 2002) and thus the species-specific
196 detrimental effects of ALAN may lead to broader cross-realm consequences. (i) Migrating
197 adult salmon serve as a food resource for terrestrial wildlife as they travel upstream to spawn.
198 Bears alone move up to 90% of all salmon biomass to land, sometimes hundreds of meters
199 from their stream of origin (Reimchen 2000). Salmon-derived minerals and nutrients are
200 further spread in the terrestrial environment through bear urine and faeces as these mammals
201 move throughout the riparian and upland forests (Hilderbrand et al. 1999). Salmon also
202 support freshwater systems by providing nutrients from their carcasses following spawning

203 (Juday et al. 1932) and play an important role in the marine food-web during their migratory
204 stage to the sea (Gende et al. 2002). (ii) ALAN-associated impacts also have negative
205 consequences for the total biomass of fish surviving to the ocean-life stage: ALAN promotes
206 asynchrony in the emergence of fry, likely increasing their predation risk (Riley et al. 2013)
207 and reducing their survival (Brännäs 1995). Moreover, given the effects of artificial light on
208 smolt dispersal, adult survival is also affected (Riley et al. 2012). How much of salmon
209 biomass is currently affected by ALAN and the magnitude of such effects for other realms
210 remains unknown. Nevertheless, a study comparing 50 watersheds in British Columbia's
211 central coast in Canada showed that salmon influence nutrient loading to plants, shifting plant
212 communities toward nutrient-rich species and declines in salmon will have the largest
213 ecological effects on smaller and less productive streams (Hocking and Reynolds
214 2011). Salmon populations are declining in many parts of the world due to a wide range of
215 anthropogenic activities (e.g. Collins et al. 2015, Falcón et al. 2020). Management actions,
216 however, rarely consider light pollution as a mitigating factor, and even fewer address cross
217 realm impacts. This is of concern given its cross-realm life history; efforts to mitigate the
218 impacts of ALAN on salmon that are solely focused in one realm may be ineffective and
219 economically wasteful if impacts from/in other realms are not considered.

220

221



222

223 Figure 1 - Schematic figure showing the potential cross-realms impacts of ALAN due to

224 effects on different life stages in salmon species. (A) Salmon spend their juvenile phase in

225 rivers before migrating to sea to grow and mature. To complete their life cycle they must

226 return to the river to spawn. (B) ALAN at sea alters vertical movement of fish resulting in a

227 mismatch between preferred light levels and optimal feeding zones. Additionally, ALAN

228 results in increased predation of fish at sea and hence a decrease in adults returning to rivers.

229 (C) ALAN along rivers disrupts synchronous emergence of juveniles resulting in increased

230 predation which then reduces the recruitment of smolts out to sea. This reduction in adults

231 returning to rivers and smolts migrating to sea results in trophic effects in both realms. (D)

232 Illustrates one trophic effect in the terrestrial environment with reduced food resources for

233 bears resulting in reduced nutrients into the terrestrial environment. Image created with

234 BioRender.com.

235

236 *Case study 2 - Aquatic insects (with terrestrial adults)*

237 Demonstrated ALAN impacts

238 Dragonflies, mayflies and mosquitoes are classic examples of secondarily aquatic
239 insects - those with an aquatic egg and juvenile phase and a terrestrial adult phase. The
240 transition from the (often protracted) juvenile aquatic environment to the terrestrial adult
241 environment is varied and taxon-specific. For example, prior to their final moult, dragonfly
242 nymphs typically move up out of the water (usually at night) onto a branch or other structure
243 where they eclose and emerge as air-breathing terrestrial adults. Mosquitoes remain in the
244 aquatic environment emerging directly into the terrestrial environment as adults, typically
245 remaining at the surface to allow their wings to dry and harden. Mayflies are hemi-
246 metabolous and thus do not have a pupal stage; instead, they emerge into the terrestrial
247 environment as a winged subadult (or sub-imago) and then rapidly moult to adults.

248 The effect of variation in moonlight on adult insect activity has been long documented
249 (Williams and Singh 1951) and it is well recognised that artificial lighting is attractive to
250 many adult insects – the behaviour is commonly exploited when trapping potential pests
251 (Shimoda and Honda 2013). Recent evidence suggests sources of ALAN (such as
252 streetlights) close to streams or water bodies may similarly change insect dispersal patterns
253 (geographic or temporal; Manfrin et al. 2017) and/or act as ecological traps for newly
254 eclosing adults (Eisenbeis et al. 2006, Perkin et al. 2011). ALAN sources can also draw
255 individuals away from the aquatic environment, an essential resource required for mating and
256 egg laying (Eisenbeis et al. 2006, Perkin et al. 2011), into suboptimal environments where the
257 risk of mortality is increased (Davies et al. 2012). Some species (e.g. dragonflies, mayflies
258 and caddisflies) are also positively polarotactic, using horizontally polarized light to locate
259 suitable water bodies for mating and egg laying (Kriska et al. 2009). In areas with
260 anthropogenic sources of polarised light (reflected off asphalt surfaces, vertical glass and

261 even vehicles), adult polarotactic behaviour can result in adults aggregating and females
262 ovipositing on suboptimal non-aquatic surfaces leading to reduced or no juvenile survival
263 (Horváth et al. 2014). Moreover, anthropogenic sources of polarised light at night can also
264 attract predatory insectivores, such as birds, lizards or spiders, resulting in increased adult
265 insect mortality (Robertson et al. 2010, Szaz et al. 2015).

266 Even when eggs are laid in an appropriate body of water, the protracted aquatic
267 juvenile phase may be vulnerable in the presence of ALAN. Evidence from other insects
268 suggests aquatic juveniles may be directly attracted to external light sources, leading to shifts
269 in foraging and other activity patterns (Kühne et al. 2021) and possible increases in predation
270 risk (Manfrin et al. 2018). Moreover, experimental evidence from terrestrial invertebrates
271 suggests prolonged exposure to ALAN during the protracted juvenile phase may influence
272 growth, development and survival as adults (McLay et al. 2017, Durrant et al. 2018, Willmott
273 et al. 2018).

274 Potential cross-realms effects and knowledge gaps

275 Secondly aquatic insects are proposed as ideal bioindicators to assess the impact of
276 cross-realm (aquatic and terrestrial) environmental change due to their sensitivity to
277 anthropogenic stressors (Villalobos-Jimenez et al. 2016). However, we lack direct evidence
278 to confirm how impacts from one realm may influence the other. Moreover, there is
279 surprisingly little information regarding the specific impact of ALAN on the independent life
280 history stages of secondarily aquatic insects: in the largest review of urban impacts on
281 dragonflies, ALAN was not even included (Villalobos-Jimenez et al. 2016). For instance, the
282 presence of anthropogenic sources of light are known to reduce reproductive success and
283 increase predation rates of many secondarily aquatic insects (as stated above), but the degree
284 to which exposure to ALAN results in selection of particular juvenile phenotypes that survive
285 to the adult stage is unknown (Hopkins et al. 2018). Ultimately, although many knowledge

286 gaps exist, such insects form a large proportion of biomass and if ALAN affects their growth,
287 survival, and distribution, this is likely to have highly problematic outcomes that span
288 multiple realms.

289

290 ***2) Impacts on species interactions that involve two or more realms***

291 The loss of, or changes in, species within a system can affect an entire cross-realm
292 network, through altered competition and/or food-web interactions, with unpredictable
293 consequences for communities, ecosystems (Eklöf and Ebenman 2006) and other, connected,
294 realms (Bugnot et al. 2019). Below, we highlight two case studies where observed or inferred
295 effects of ALAN for one species or group are expected to affect multiple realms through
296 species interactions and knock-on effects.

297

298 ***Case study 3 - Fishing bats: terrestrial mammals specialised for feeding in aquatic*** 299 ***ecosystems***

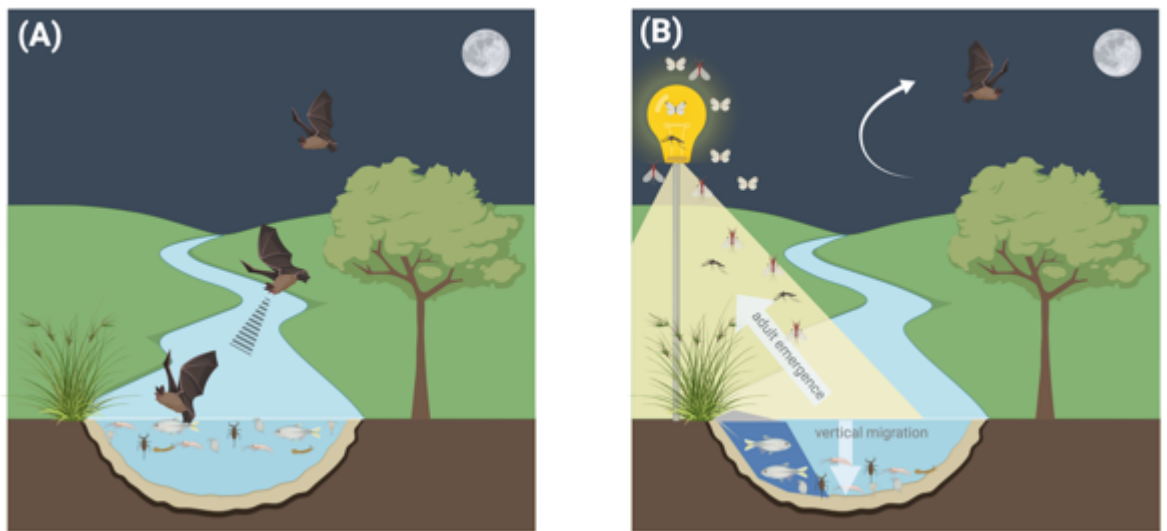
300 *Demonstrated ALAN impacts*

301 Worldwide, there are 16 species of fishing or trawling bats (e.g. from the genus
302 *Myotis*). This group has ecological and foraging specialisations that make them reliant on
303 both terrestrial and aquatic realms (Campbell 2011). Fishing bats roost diurnally in caves,
304 aqueducts, bridges, tunnels and tree cavities in the vicinity of water sources (Campbell 2009,
305 Gorecki et al. 2020) and forage exclusively nocturnally on aquatic prey using their feet to
306 trawl the surface of water for fish and aquatic insects (Dwyer 1970, Law and Urquhart 2000,
307 Campbell 2007). Neither group of bats can detect submerged prey (Suthers 1965) and instead
308 rely on echolocation of water surface irregularities created by fish and aquatic invertebrates
309 (Thompson and Fenton 1982).

310 ALAN has direct and indirect effects on the bat communities. Of primary concern is
311 the fact that fishing bats are largely light averse and thus either actively avoid lit areas,
312 possibly due to increased risk of predation (Straka et al. 2016), and/or reduce their feeding
313 attempts when waterways are lit (Haddock 2019). Indirectly, light affects prey abundance.
314 Aerial invertebrates are attracted to sources of ALAN, but fishing and trawling bats are
315 unable to capitalise on this increased abundance due to their own aversion to light. Coupled
316 with this, many aquatic invertebrates (potential prey items for bats) exhibit diel vertical
317 migration: moving downwards from the water's surface to deeper water during the day and
318 moving upwards to the surface during the night (Perkin et al. 2011, Mehner 2012) where they
319 forage or potentially emerge as adult aerial invertebrates from the aquatic realm (Manfrin et
320 al. 2017). In areas exposed to ALAN, nocturnal vertical migration of invertebrates to the
321 surface is reduced and fewer adults eclose resulting in fewer opportunities for fishing bats to
322 forage for prey.

323 Potential cross-realms effects and knowledge gaps

324 Light sources near aquatic habitats can therefore impact bats through impacts on their
325 ability to forage. Again, the full consequence of this impact to populations is unknown, and
326 even smaller is our understanding on how population declines can in turn influence, more
327 generally, their terrestrial habitats. However, due to the known association of these animals to
328 both aquatic and terrestrial realms, knock-on effects on both are expected. Conservation and
329 management efforts should thus include ALAN as a potential threat for these highly
330 specialised species.



331

332 Figure 2 - (A) Schematic figure depicting the aquatic ecosystem with fishing bats under
 333 natural light (B) and how artificial light at night influences prey species. As artificial light is
 334 introduced, aquatic prey species migrate into shadows, sediment or to greater depths, making
 335 them unavailable to bats. Additionally, some aquatic insects emerge as aerial adult forms that
 336 are attracted to light. Fishing bats avoid lit areas and cannot switch foraging strategies to take
 337 advantage of the new aerial prey that is attracted to lights. Image created with
 338 BioRender.com.

339

340 ***Case study 4 – Shifting energy flows between realms via impacts on orb-web spiders and***
 341 ***aquatic insect communities***

342 ***Demonstrated ALAN impacts***

343 In riparian zones, increased predation pressure on emerging aquatic insects around
 344 ALAN through the attraction to nocturnal lighting by both predators and prey can reduce the
 345 transfer of biomass from aquatic to terrestrial systems. Short-term (two-month) exposure to
 346 ALAN was linked to an increased abundance and associated body mass of riparian long-
 347 jawed orb weavers (family Tetragnathidae) (Parkinson et al. 2020). These effects were more

348 pronounced for females compared to males and were concordant with greater numbers of
349 prey items captured in spider webs under ALAN compared to webs under natural night-time
350 conditions. However, a comparable, but longer-term, study (one year) found that although
351 spider density initially increased (as in the previous study), there was a long-term decrease in
352 spider density, as well as a decrease in the emergence of aquatic insects (Meyer et al. 2013).
353 ALAN therefore shifts biomass from dark areas into artificially illuminated areas and
354 dramatically shifts the distribution, overall abundance, and diversity of insect communities,
355 reducing their abundance as prey for predators (Perkin et al. 2014, Manfrin et al. 2017,
356 Parkinson et al. 2020).

357 Potential cross-realms effects and knowledge gaps

358 By altering both the abundance and predation success of terrestrial predators, as well
359 as the distribution and abundance of aquatic prey, ALAN can drive shifts in predator-prey
360 interactions across realm boundaries, altering flows of energy between aquatic and terrestrial
361 systems, with important consequences for both realms. Resource exchange from terrestrial to
362 aquatic realms is an intrinsic facet of riparian habitats (Baxter et al. 2005). Spiders are
363 important predators in riparian zones and can obtain more than 50% of their nutrition from
364 aquatic sources, especially insects (Collier et al. 2002). Therefore, effects of ALAN on the
365 diversity, abundance and distribution of spiders (both free-living and web-building), and/or
366 the community of aquatic insects in riparian zones can alter cross-realm fluxes, with
367 important regional and global implications for both terrestrial and aquatic realms (Manfrin et
368 al. 2017). The consequences of these effects of ALAN depend on the time-scale considered
369 and may be sex-specific.

370

371

372 **3) Impacts on transition zones**

373 In areas where light pollution affects critical transition zones (e.g. at ecosystem
374 boundaries, affecting two or more realms) it is likely there will be consequences for
375 ecosystems and function and service. Furthermore, transition zones tend to be
376 disproportionately affected by ALAN, since many urban settings, where ALAN is prevalent,
377 are developed near waterways (Kummu et al. 2011). Estuaries and coastal wetlands are
378 critical transition zones that link freshwater habitats with marine and terrestrial environments
379 (Levin et al. 2001). These zones perform important ecological functions such as nutrient
380 cycling and regulation of water and nutrient fluxes between realms (Levin et al. 2001).

381 Natural light at the air-water interface is a key factor linking terrestrial and aquatic
382 realms. The amount of light that reaches the water surface in freshwater or coastal systems
383 depends on the surrounding terrestrial habitat: structurally complex terrestrial environments,
384 such as forested riparian zones, reduce the amount and colour of light reaching the water
385 surface (Endler 1993). Species also vary extensively in their sensitivities to multiple light
386 properties (Gaston et al. 2012, Land and Nilsson 2012), and transition zones support several
387 specialised species that have adapted to these complex lighting environments. For example,
388 in estuaries with turbid waters, high loads of suspended material and low ambient light levels,
389 fish species, such as the flathead grey mullet (*Mugil cephalus*), have evolved morphological
390 traits that support dim-light (i.e. scotopic) vision, such as high rod density in the retina
391 (Zapata et al. 2019). Similarly, the freshwater three-spine stickleback (*Gasterosteus*
392 *aculeatus*) has a highly specialised visual sensitivity important for mate selection in both
393 clear versus tannin-stained lakes (Boughman 2001). Transition zones, therefore, are
394 significant sites for understanding and managing cross-realm impacts of ALAN, both due to
395 the vulnerability of organisms inhabiting these zones, and the prevalence of light pollution
396 near waterways.

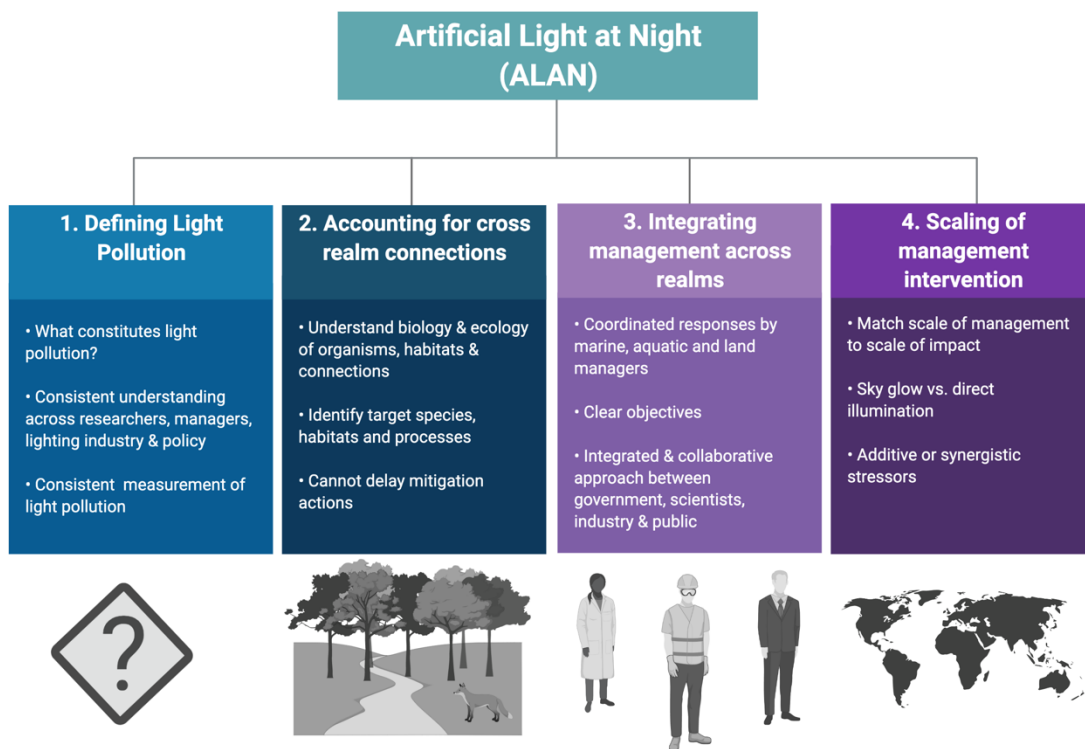
397 Shifts in the flow of resources in riparian zones – the interface between land and
398 rivers or streams – can have impacts across multiple realms (see Case study 4). In their recent
399 comprehensive review, Zapata et al. (2019) outlined a multitude of ways ALAN can
400 specifically affect estuaries and highlighted potential cross-realm implications. For example,
401 ALAN-induced delays in the leaf fall of deciduous trees (Bennie et al. 2016) can in turn
402 reduce the input of nutrients from leaf detritus into aquatic systems, causing potential shifts in
403 the biogeochemistry of aquatic systems (Zapata et al. 2019). Furthermore, in their review,
404 Falcón et al. (2020) discussed ALAN effects on riparian ecosystems and Sullivan et al.
405 (2019) recently demonstrated the impacts of ALAN on riparian systems through shifts in the
406 community structure of invertebrates, consequently altering the flows of energy between
407 aquatic and terrestrial systems. Given these direct examples and published review of the
408 impacts of ALAN on transition zones and flow-on effects across realms, we have not
409 provided case studies here to further illustrate this mechanism. Instead, we want to highlight
410 the importance of prioritising transition zones for management actions to limit the impacts of
411 light pollution across multiple realms.

412

413 **CHALLENGES AND PRACTICAL SOLUTIONS FOR RESEARCH AND** 414 **MANAGEMENT OF ALAN**

415 Several challenges exist that need to be addressed for the impacts of light pollution to
416 be effectively understood and managed, both within and across realms. A major difficulty
417 (and potential point of contention) encountered when dealing with cross-realm issues is
418 determining the boundaries for management and governance (Pittman and Armitage 2016).
419 For example, land-based sources of ALAN may indirectly influence the productivity of
420 aquatic systems through its impact on nutrient inputs from terrestrial sources through e.g.
421 changes in the leaf fall patterns of deciduous trees. In this case, areas are separated by

422 physical and jurisdictional boundaries (e.g. land and coastal managers) and potentially social
 423 boundaries (different communities or social networks). Here, we propose a framework for
 424 cross-realm management, which builds on previous frameworks for conservation and
 425 management across-realms (e.g. Beger et al. 2010, Alvarez-Romero et al. 2015, Giakoumi et
 426 al. 2019, Threlfall et al. 2021), but with a specific focus on light pollution (Figure 3).



427
 428 Figure 3 - Proposed framework to explore the cross-realm impact of artificial light at night. 1.
 429 Defining light pollution – requires a shared understanding of what constitutes light pollution,
 430 and that its meaning and measurement is consistent across all stakeholders; 2. Accounting for
 431 cross-realm connections – requires knowledge of the ecological environment, the organisms,
 432 target species and how cross realm impacts intersect; 3. Integrating effective cross-realm
 433 management – requires all stakeholders to be clear on objectives and outcomes; and 4.
 434 Effective scaling of management integrations – requires the scale of the management
 435 intervention to match the scale of impact. Image created with BioRender.com.

436

437 ***Challenges and practical solutions***

438 *1 - Defining light pollution*

439 One of the main challenges for driving practical solutions to manage ALAN is
440 agreeing to a collective understanding of how and when lighting should be defined as
441 pollution (Schulte-Römer et al. 2019). Here, we define light pollution as light introduced into
442 the environment by humans at intensities that are higher than the natural level at that time for
443 the given environment and that has the potential to cause harm to humans and/or the
444 environment. In a recent analysis, Schulte-Römer et al. (2019) found that light pollution
445 experts (including scientists and managers) had a stronger and more consistent view of what
446 constitutes light pollution than lighting professionals (such as lighting designers, urban
447 planners and engineers). Importantly, however, both groups had very skewed views when
448 considering potential issues caused by light in areas where it is ‘unwanted’, depending on the
449 habitat or realm. Of the respondents identified as light pollution experts (n = 89),
450 approximately 90% considered light to be pollution when it obscures the visibility of stars, or
451 when fixtures were installed close to observatories. In contrast, only 66% of experts surveyed
452 considered lighting as pollution when it was installed close to bodies of water. Among the
453 respondents identified as lighting professionals (total of n = 67 respondents), this dropped to
454 only 17%. These results highlight a common misconception, and a massive global problem,
455 namely, that light is a ‘land’ problem rather than of fundamental significance for all
456 ecosystems on earth. These findings also ignore the critical need for fluctuating light levels
457 (both day and night) that have characterised the evolutionary history of that life. Therefore,
458 the first steps to successfully managing light pollution within and across realms are to (i)
459 raise awareness of the importance of fluctuating light regimes for ecological process; (ii)
460 enhance understanding of the impacts of artificial light across all realms: terrestrial,

461 freshwater and marine environments; (iii) broaden knowledge regarding the impact that light
462 within one realm can have for biodiversity and ecosystem function within other realms and
463 (iv) understand the ‘acceptable’ levels of ALAN for both the local ecological communities
464 and society (i.e. trade-offs between ecological impacts and societal needs or desires).
465 Critically, this needs to include multiple stakeholders, including the general public.

466

467 *2 - Accounting for cross-realm connections*

468 The next step in managing light pollution across realms is to understand the biology
469 and ecology of organisms and habitats of interest and their potential linkages, so that
470 management interventions can more fully account for connections across realms. Ideally, the
471 extent of the impact of ALAN on target individuals, populations, habitats and systems, as
472 well as the mechanisms driving these changes, will be well-known within and across realms.
473 However, we acknowledge that, unfortunately, the current state of habitat degradation
474 worldwide and rapid expansion of ALAN means that we cannot afford delaying mitigation
475 actions until the impacts, or even the potential unintended risks of management interventions,
476 are fully understood (Mayer-Pinto et al. 2019). Therefore, we need to keep gathering the -
477 still much needed - scientific information on the effects of ALAN, within and across realms,
478 while, at the same time, implementing local, regional and global best practice guidelines to
479 prevent or lessen such impacts.

480

481 *3 - Integrating management across realms*

482 A key challenge associated with managing the impact of ALAN across realms is the
483 lack of collaboration between different stakeholders and the existence of methodological
484 disparities across realms. The compartmentalisation that can exist within governance
485 structures, such as within and between local, state/territory and federal government agencies

486 inevitably generates a lack of consistency in management decisions which are exacerbated
487 when considerations involve multiple realms (and thus multiple stakeholders). Contributing
488 factors include poor communication, differing and potentially competing priorities and a lack
489 of collaboration among the sectors and agencies responsible for planning and environmental
490 protection in the different realms; a lack of spatial data on cross-realm processes; and, logistic
491 difficulties associated with adapting existing decision-tools and coordinating different
492 governance systems to fit the current purpose (Alvarez-Romero et al. 2015 and references
493 therein).

494 To successfully implement cross-realm management strategies, some key general
495 steps (adopted and modified from e.g. Alvarez-Romero et al. (2011), Bugnot et al. (2019),
496 Threlfall et al. (2021)) can be taken. First and foremost, a clear objective regarding the
497 desired outcomes is necessary. For issues pertaining to light pollution, these can include
498 minimising or eliminating the effects of ALAN on ecologically, culturally and/or
499 commercially important target species/groups or a target area (e.g. a transition zone,
500 migratory pathways or a protected area). This necessitates an integrated and collaborative
501 approach with policy makers, regulators, scientists, lighting designers, developers and the
502 general community, including First Nations People, to identify potential conflicting interests
503 and devise solutions accordingly. Ultimately, we need to both unify terminologies and agree
504 on desired outcomes (Webb 2012, Bugnot et al. 2019), and, ideally, understand potential
505 thresholds of ‘acceptable’ artificial light levels across different species and realms, which will
506 likely involve a compromise between levels of ecological impacts caused by ALAN and
507 societal needs or desires.

508 Determining ALAN thresholds, however, requires standardised measurements of light
509 *per se*. Currently, there is great inconsistency in instrumentation and light parameters within
510 and across realms. Discrepancies in lighting measurements exist for valid and practical

511 reasons – e.g. the measurement and instrument used needs to match the scale of both the light
512 pollution being measured (i.e. direct sources of light vs skyglow) and the ecological or
513 biological response of interest (e.g. insect attraction to a street light vs bird migration).
514 Moreover, as far as we know, there is not yet available affordable and easy-to-use
515 instrumentation to adequately measure light levels under water. However, there is a clear and
516 urgent need to standardise, where possible, the measurement of light pollution, so that
517 outcomes are comparable and applicable across realms (see Box 3 for further discussion). It
518 is important to note, however, that knowing relevant light ‘levels’ is not enough for effective
519 management for ecological outcomes. At the extreme, any light that is not natural in its origin
520 is likely to interfere with ecological process. Thus, perhaps of greater importance, we need to
521 be able to measure and understand how light properties (including spectra and intensity)
522 affect organisms and habitats in multiple realms. Standardising how and which properties of
523 light are measured will facilitate communication of clear and specific recommendations
524 (including biologically relevant thresholds) between researchers, practitioners and managers.
525 This will permit informed decision making when considering potential impacts across
526 different habitats and realms and allow better assessment of the risks when night-time
527 illumination is unavoidable and/or socially desirable.

528

529 *4 - Scaling of management intervention*

530 Ultimately, there is a need to match the scale of the management intervention to the
531 scale of impact (Threlfall et al. 2021). Light pollution impacts occur at the landscape scale,
532 and include impacts caused by sky glow, light scattered in the atmosphere (Cinzano et al.
533 2001, Falchi et al. 2016), and those caused by direct illuminance from light sources (e.g.
534 streetlights). Impacts caused by direct illuminance are, in theory, easier to mitigate, than
535 impacts caused by sky glow – which can be an issue even tens (and possibly hundreds) of

536 kilometres from urban light sources (Gaston et al. 2012) and require management
537 interventions at much larger, landscape level, scales to prevent or mitigate cross-realm
538 impacts. For example, research has shown that light pollution can spill into otherwise
539 protected areas up to 15 km from urban centres (McNaughton et al. 2021). Additionally, a
540 recent study has highlighted the potential for synergistic interactions between sky glow and
541 direct illuminance (Dickerson et al, unpublished data). Management actions therefore need to
542 consider, whenever possible, multiple spatial scales to mitigate light pollution and avoid
543 cross-realm impacts. Extensive examples on specific interventions and management
544 strategies can be found in the literature (Gaston et al. 2012, DAWE 2020).

545 Light pollution is just one of a multitude of anthropogenic stressors associated with
546 urbanisation (Dominoni et al. 2020), which can also cross realm boundaries. Therefore,
547 management interventions should also consider potential additive or interacting impacts from
548 multiple stressors (Hale et al. 2017). For example, ALAN and night-time warming have non-
549 additive interactive effects on the predation of aphids by lady beetles, decreasing aphid
550 population densities (Miller et al. 2017). Similarly, particular traits in birds can be impacted
551 by both ALAN and noise pollution: light pollution is associated with advancement in
552 reproductive phenology of several species of birds while noise decreased clutch size of
553 closed-habitat (i.e. forests) birds (Senzaki et al. 2020). Interactive effects of anthropogenic
554 stressors with ALAN remain, however, poorly understood (Falcón et al. 2020).

555 Understanding, or at a minimum identifying, other stressors that may interact with or act
556 simultaneously with ALAN will enhance cross-realm management outcomes. Moreover,
557 climate change adds additional challenges to cross-realms studies since it increasingly
558 modifies key land-sea ecological and social processes, therefore increasing the urgency for
559 transboundary management initiatives.

560

561 **CROSS-REALMS MANAGEMENT SUCCESS**

562 There have been few examples of successful management of ALAN which have
563 resulted in a reduction of cross-realm impacts, and most of these examples involved
564 management interventions that targeted a single species rather than an assessment at
565 community or ecosystem levels. Successful examples include: 1) the mitigation of impacts on
566 shearwaters (Phillip Island, Victoria, Australia) through changes to the timing and colour of
567 street lights, particularly during critical periods of the life cycle – i.e. fledging (Rodríguez et
568 al. 2014, Rodríguez et al. 2017); and 2) legislation related to nesting marine turtles (DAWE
569 2020). Below, we expand on the latter.

570 Marine turtles have complex life histories that cross marine and terrestrial realms, and
571 are considered key indicators of ecosystem health (Haywood et al. 2019). Light pollution can
572 reduce the reproductive viability of turtle stocks by disrupting critical behaviour such as the
573 ability of hatchling marine turtles to successfully reach the ocean (Witherington and Bjorndal
574 1991). Light in nearshore waters (e.g. boats on anchor, jetties, or coastal lighting) can
575 influence the offshore dispersal of hatchlings in the critical minutes and hours after they leave
576 the beach. Attraction to artificial lights increases the time hatchlings spend crossing predator
577 rich nearshore waters before reaching the safety of deep water offshore, thus increasing their
578 vulnerability to predation (Harewood and Horrocks 2008, Thums et al. 2016, Wilson et al.
579 2018); and as predators are also attracted to the same lights, predation pressure can be high.
580 In Australia, activities that involve artificial light at night that is likely to impact marine
581 turtles must be referred for environmental assessment. Proponents must demonstrate, via
582 formal risk assessments, how the impact of ALAN on all age classes of marine turtles will be
583 mitigated and adaptively managed. Mitigation measures that benefit marine turtles have been
584 summarised in the National Light Pollution Guidelines for Wildlife Including Marine Turtles,
585 Seabirds and Migratory Shorebirds (DAWE 2020) and include 1) management of the

586 physical aspects of the light, such as intensity (lumen output), colour (wavelength) and
587 elevation above dark horizons behind the beach, 2) the maintenance of dark zones between
588 turtle nesting beaches and light sources, and 3) shielding and targeting of light fixtures to
589 avoid direct visibility and limiting sky glow (DAWE 2020). Given light pollution sources
590 that can affect turtles can be both marine and terrestrial, management actions in both realms
591 are likely required, with the collaboration of terrestrial and aquatic ecologists and lighting
592 professionals, to successfully avoid terrestrial-aquatic impacts.

593 It is important to note, however, that, even though management actions outlined here
594 were focused on one particular group of organisms (e.g. marine turtles), a general
595 understanding of both the terrestrial and marine realms and potential linkages among them, as
596 well as a clear desired outcome, was necessary to devise efficient strategies. None of which,
597 could have been achieved without collaboration among different stakeholders in each
598 individual realm.

599

600 **MITIGATING IMPACTS OF FUTURE LIGHTING**

601 There is increasing recognition that conservation and management strategies should be
602 designed to account for cross-realm connections (e.g. Threlfall et al. 2021, Tulloch et al.
603 2021). A recent study developed a national-scale conservation framework that incorporated
604 linkages among the marine, freshwater and terrestrial realms, to select protected areas for
605 minimising the threats of both land-use and climate change (Tulloch et al. 2021). The cross-
606 realm approach resulted in changes to both terrestrial and marine priorities compared to when
607 connections among realms were not considered. The authors also argued that a cross-realm
608 approach allowed the identification of potential trade-offs and opportunity costs of
609 conservation versus ecological benefits, as well as the implementation of interventions with

610 multiple objectives (such as habitat management and biodiversity protection) (Tulloch et al.
611 2021).

612 Increasing the uptake of a cross-realm management approach requires increased and
613 improved communication between researchers, lighting practitioners, managers and
614 regulators that work within and across different realms. The creation of professional networks
615 is a great way to begin such conversations. In Australia, the Network for Ecological Research
616 on Artificial Light (NERAL; www.neralaus.com) was established to provide a platform to
617 connect researchers and practitioners working towards mitigating the impacts of light
618 pollution within and across realms. NERAL is a professional network of academic scientists
619 and consultants, with a wide range of expertise, including terrestrial and marine ecologists
620 and physiologists, and managers from local and federal government agencies. A primary aim
621 of the network is to increase communication between scientists and managers working on
622 different species, habitats and/or realms. This will allow: 1) managers to easily access
623 information crucial to developing and implementing interventions to prevent or mitigate light
624 pollution impacts, and 2) researchers to identify management priorities and provide evidence-
625 based information to shape management interventions. Networks that have a strong multi-
626 realm focus such as NERAL are important, as they enable a more holistic understanding of
627 issues related to ALAN. They can also provide an opportunity to develop standardised
628 methods for measuring light so that the impacts can be compared across realms. This holistic
629 approach can then be translated into the ongoing implementation of strategies to reduce
630 impacts of ALAN across terrestrial, marine and freshwater realms.

631

632 **BOX 1) LIGHT AS AN ECOLOGICAL TRAP**

633 Ecological traps arise when animals are attracted to and remain in poor-quality
634 habitats where their fitness is compromised (Hale and Swearer 2016). ALAN can cause
635 ecological traps by influencing both the habitat selection decisions of animals and their
636 fitness consequences. The orb-web spiders and aquatic insect community case study
637 presented here clearly illustrates this – the adult stages of aquatic insects are attracted to
638 artificial light where they suffer higher mortality because of the high density of webs. This
639 case study provides further evidence of how ecological traps caused by ALAN can impact on
640 cross-realm linkages. In this case, ALAN strengthens the magnitude of cross-realm predator-
641 prey interactions. Specifically, the higher attraction and mortality of aquatic insects leads to
642 increased aquatic-to-terrestrial subsidy flux (e.g. Manfrin et al. 2017).

643 Artificial light can also interfere with the migratory behaviour of species that occupy
644 different realms as part of their life cycle. A well-known example of this is the impact of
645 ALAN on the dispersal behaviour of sea-turtle hatchlings. Nocturnally emerging hatchlings
646 are attracted to artificial lighting from coastal development. Crawling towards an artificial
647 light source can result in predation (Erb and Wyneken 2019), impair their ability to swim
648 offshore (Lorne and Salmon 2007), leading to reduced rates of offshore migration and rates
649 of transition between life stages (Wilson et al. 2019).

650 Lastly, ALAN could increase cross-realm rates of disease transmission due to its
651 impact on vector biology, such as biting mosquitoes. For example, in a recent study by Fyie
652 et al. (2021), artificial light masked natural daylength change which is the trigger for
653 diapause, meaning mosquitos remained reproductively active for longer and produced more
654 aquatic larvae. ALAN exposed mosquitos also had increased rates of blood feeding compared
655 to control mosquitos. Given the preference for humans to associate with artificially lit

656 environments at night, this suggests both changes in human and vector behaviour have
657 resulted in a largely unrecognized ecological trap for humans.

658

659 **BOX 2) CROSS-REALM EXPLOITATION OF RESOURCES USING ARTIFICIAL**
660 **LIGHT AT NIGHT**

661 Artificial light at night is known to attract and/or aggregate many organisms. This
662 effect can be exploited by predator species within and across realms, if, for example, a
663 terrestrial predator is exploiting an aggregation of aquatic organisms to a light source. One of
664 the best cross-realm examples of how ALAN can be used to exploit resources is the use of
665 artificial light by humans during night-time fishing.

666 The attraction of many fish and aquatic invertebrates to light has been known for
667 thousands of years, and artificial light has been used by humans to improve fishing efficacy
668 for centuries (Yami 1976). Light at night attracts small fish, insects and/or plankton through
669 positive phototaxis, disorientation, or curiosity (Marchesan et al. 2005), which in turn attracts
670 larger predatory fishes and invertebrates (Becker et al. 2013). Historically, humans exploited
671 this behaviour by lighting a fire on a beach to attract fish into the shallows to facilitate
672 harvest (e.g. by spearing or netting) (Yami 1976). Today, incandescent, fluorescent, metal
673 halide, and LED above-water and underwater lights are used for artisanal and industrialized
674 fishing practices worldwide to increase harvest (Solomon and Ahmed 2016, Nguyen and
675 Winger 2019). In fact, certain fisheries cannot operate effectively without the use of lights,
676 such as the squid jigging fishery. Jigging for squid dates back to antiquity in many parts of
677 the world, however in the recent century, the addition of artificial light to jigging gear has
678 substantially increased landings due to the effect of light at night on attracting and
679 concentrating squid (Solomon and Ahmed 2016).

680 The effects of ALAN on fish attraction/aggregation are not lost on recreational
681 fishers; recreational fishers often target artificially lit areas for night fishing, as they know
682 certain target game species will follow baitfish into the illuminated areas (Cooke et al. 2017).
683 Urbanization has led to an increase in artificial light installations in coastal areas,
684 illuminating a substantial portion of shallow aquatic habitats at night (Davies et al. 2014,
685 Davies et al. 2016), and has therefore created ample opportunities for recreational fishers to
686 exploit artificial lighting (i.e. light pollution) to increase catch rates.

687 The increased harvest resulting from fishing practices using ALAN can lead to
688 overfishing and increased rates of bycatch in a fishery which may can have negative impacts
689 on fished populations (e.g. reduction in size and altered life-history traits) (Solomon and
690 Ahmed 2016) and thus ecological consequences for the marine or freshwater realms (e.g.
691 through trophic cascades). However, since responses to ALAN are species-specific, ALAN
692 can be used by humans to both increase fishing harvest and reduce catch rates of different
693 species. The use of artificial light has been recognized as a potential tool for bycatch
694 reduction in commercial fisheries, and therefore ALAN can also be exploited to mitigate
695 cross-realm impacts through minimizing effects of fishing on non-target organisms. Research
696 on the use of artificial light to reduce bycatch has demonstrated varying levels of success
697 (e.g. Hannah et al. 2015, Larsen et al. 2018, Lomeli et al. 2018) and is dependent on species
698 of interest, light properties tested, and proper placement/location of (often LED) lights within
699 the fishing gear. However, the use of artificial light to deter adult sea turtles has also proved
700 to be effective (e.g. Wang et al. 2010, Virgili et al. 2018) resulting in LED lights now widely
701 applied worldwide in pelagic gillnet fisheries to reduce sea turtle bycatch (Nguyen and
702 Winger 2019). This positive use of artificial light demonstrates that with species-specific
703 knowledge, it is possible to harness the effects of ALAN for positive impacts across realms.

704

705 **BOX 3: DISCREPANCIES IN LIGHT MEASUREMENTS**

706 A complicating factor influencing the ability of scientists to confidently predict the
707 impact of light on a sensitive receptor is the lack of an agreed upon standard method for
708 modelling, measuring and monitoring light or skyglow (e.g. Jechow and Hölker 2019,
709 Jechow et al. 2019, Kalinkat et al. 2021). Instrument types and applications vary widely:
710 instruments include lux meters, spectrometers, and cameras which measure light emitted
711 directly from a source or light reflected from a surface, from overhead looking down on the
712 earth (satellite based) or from the ground looking up or horizontally across the landscape.
713 Limitations include: restrictions in the wavelengths they measure (i.e. they do not measure all
714 wavelengths across the entire visible spectrum), detection limits that are not low enough to
715 measure sky glow or intensities that elicit a biological response, highly technical instruments
716 requiring specialised knowledge to operate and maintain, and a wide range of different
717 measurement units.

718 Arguably, many of the existent ‘disparities’ arise because different instruments are
719 designed to measure different things, depending on the objectives of the users. For example,
720 studies aiming to measure large-scale environmental effects due to sky glow will (and
721 should) measure different variables (and consequently use different instruments) than studies
722 which the primary aim is to evaluate the effects of street-light on one species of insect.
723 Nevertheless, whenever possible, studies with similar objectives and/or operating at similar
724 spatial scales, should try to standardise measurements. Crucially it is important to understand
725 the operating limits of even the simplest instruments, as instruments can be misused or used
726 for an inappropriate environment (Longcore et al. 2020). Similarly, the literature
727 acknowledges that there are no conclusive intensity thresholds below which artificial light is
728 not harmful to species and habitats (Schroer et al. 2020), and even the low intensity light
729 characteristic of skyglow can affect organisms (Grubisic et al. 2019, Kupprat et al. 2020).

730 Attempts to compare or standardise measurements across realms adds further
731 complications. For instance, while remote sensing techniques are commonly used as a best
732 proxy to quantify the amount of artificial light at night on terrestrial systems, there are serious
733 challenges associated with the use of this technology in water bodies/underwater (see the
734 extensive discussion in Jechow and Hölker 2019). Furthermore, different disciplines often
735 use different physical quantities and units for measuring light, creating confusion even among
736 experts (Jechow and Hölker 2019). For instance, much of the existing data on the quantity
737 and quality of light reaching both terrestrial and aquatic systems assess different physical
738 parameters (spectral irradiance, illuminance); have used several different instruments to
739 acquire measurements (e.g. SQM, luxmeter, spectrometer, digital camera); and, report
740 outcomes using different measurement units (lux, candela, magnitudes, Watts). Therefore, as
741 stated by Jechow and Hölker (2019), ‘there is no clear coherence between these
742 measurements, although each of them was well designed and conducted’. Cross-realm
743 assessment and management of light pollution is impeded by the discrepancies in
744 measurements of light pollution across systems and disciplines. However, standardization of
745 measurements across species level responses, systems, and realms of interest is incredibly
746 challenging, as measurements currently generally differ for valid, practical reasons, such as
747 the ecological and spatial scale of interest. This challenge highlights the value of cross-realm
748 and cross-discipline networks for developing solutions that allow efficient conservation and
749 management actions across species, habitats and realms.

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754

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766

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776

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