



Article title: Light pollution: A landscape-scale issue requiring cross-realm consideration

Authors: Mariana Mayer-Pinto[1], Theresa M. Jones[2], Stephen E. Swearer[3], Kylie A. Robert[4], Damon Bolton[5], Anne E. Aulsebrook[6], Katherine A. Dafforn[7], Ashton L. Dickerson[8], Alicia M. Dimovski[9], Nikki Hubbard[10], Lucy K. McLay[11], Kellie Pendoley[12], Alistair G.B. Poore[13], Michelle Thums[14], Nikolas J. Willmott[15], Kaori Yokochi[16], Emily K. Fobert[17]

Affiliations: Centre for Marine Science and Innovation, Evolution and Ecology Research Centre, School of Biological, Earth and Environmental Science, University of New South Wales, Sydney, NSW 2052, Australia[1], School of BioSciences, University of Melbourne, Parkville, VIC 3010, Australia[2], National Centre for Coasts and Climate (NCCC), School of BioSciences, University of Melbourne, Parkville, VIC 3010, Australia[3], Department of Ecology, Environment & Evolution, La Trobe University, Melbourne, VIC 3086, Australia[4], School of BioSciences, University of Melbourne, Parkville, VIC 3010, Australia; Department of Behavioural Ecology and Evolutionary Genetics, Max Planck Institute for Ornithology, Seewiesen 82319, Germany[5], Department of Earth and Environmental Sciences, Macquarie University, North Ryde, NSW 2109, Australia[6], Agriculture Victoria Research, Bundoora, VIC 3083, Australia[7], Pendoley Environmental Pty Ltd, 12A Pitt Way, Booragoon, WA 6154, Australia[8], Australian Institute of Marine Science, Indian Ocean Marine Research Centre, University of Western Australia, Crawley, WA 6009, Australia[9], Centre for Integrative Ecology, School of Life and Environmental Sciences, Deakin University, Burwood, VIC 3125, Australia[10], School of BioSciences, University of Melbourne, Parkville, VIC 3010, Australia; College of Science and Engineering, Flinders University, Bedford Park, SA 5042, Australia[11]

Orcid ids: 0000-0001-9679-7023[1]

Contact e-mail: m.mayerpinto@unsw.edu.au

License information: This is an open access article distributed under the terms of the Creative Commons Attribution License (CC BY) 4.0 <https://creativecommons.org/licenses/by/4.0/>, which permits unrestricted use, distribution and reproduction in any medium, provided the original author and source are credited.

Preprint statement: This article is a preprint and has not been peer-reviewed, under consideration and submitted to UCL Open: Environment Preprint for open peer review.

DOI: 10.14324/111.444/000103.v1

Preprint first posted online: 25 October 2021

Keywords: ALAN, artificial light at night, light pollution, multi-disciplinary, adaptive management, ecological connectivity, Sustainability, Environmental science, People and their environment, Environmental protection, Urban studies, Sustainable and resilient cities, Biodiversity, Conservation

22nd October 2021

Professor Dan Osborn
Editor-in-Chief
UCL Open: Environment



Dear Professor Osborn,

Please find our manuscript entitled '**Light pollution: A landscape-scale issue requiring cross-realm consideration**', by M. Mayer-Pinto and collaborators, which we would like you to consider for publication in *UCL Open: Environment*.

Artificial light at night (ALAN) is expected to profoundly impact most ecosystems on the planet by disrupting a fundamental driver of and evolutionary processes: natural light cycles. ALAN is an emergent global stressor and affects approximately one-quarter of the planet. It can impact a wide range of organisms and habitats as well as multiple realms. Terrestrial (including air and land), marine, and freshwater realms are inherently linked through ecological, biogeochemical and/or physical processes. Nevertheless, current management practices for light pollution rarely consider connectivity between realms.

Here, we discuss the ways in which ALAN can have cross-realm impacts and provide case studies for each example discussed. We identify three main ways in which ALAN can affect two or more realms: 1) through impacts on species that have life cycles and/or stages on two or more realms, such as diadromous fish that cross realms during ontogenetic migrations and many terrestrial insects that have juvenile phases of the lifecycle in aquatic realms; 2) impacts on species interactions that occur across realm boundaries, and 3) impacts on transition zones or ecosystems such as mangroves and estuaries.

We consider the consequences of taking a single-realm approach to light pollution management and propose a framework for cross-realm management of ALAN, incorporating both theoretical and empirical considerations. We then discuss current challenges and potential solutions to increase the uptake of a cross-realm approach for light pollution management. Given ALAN is projected to increase in all three realms in response to continuing human population growth, cross-realm management will be critical for ensuring the ongoing resilience of ecosystems.

We believe this critical and timely article will be of broad interest to the readers of *UCL Open: Environment* and fills important gap in the research.

On behalf of the authors, I can confirm that none of the manuscript has been previously published or is being considered for publication in any other journal or a book. The authors have seen the manuscript and agree to its submission for publication.

Yours Sincerely,

A handwritten signature in black ink, appearing to read 'Mariana Mayer Pinto', written over a horizontal line.

Dr Mariana Mayer Pinto
Scientia Senior Lecturer
Centre of Marine Science and Innovation; Evolution & Ecology Research Centre
School of Biological, Earth and Environmental Sciences
University of New South Wales, Sydney NSW 2052

1 **Light pollution: A landscape-scale issue requiring cross-realm consideration**

2 Mayer-Pinto M.^{1*}; Jones T.M.²; Swearer S.E.³; Robert K.A.⁴; Bolton D.¹; Aulsebrook A.E.

3 ^{2,5}; Dafforn K.A.⁶; Dickerson A.L.²; Dimovski A.M.⁴; Hubbard N.¹; McLay L.K.⁷;

4 Pendoley, K.⁸, Poore, A.G.B.¹; Thums, M.⁹, Willmott N.J.²; Yokochi K.¹⁰ & Fobert E.K.^{2,11}

5

6 ¹ Centre for Marine Science and Innovation, Evolution and Ecology Research Centre, School
7 of Biological, Earth and Environmental Science, University of New South Wales, Sydney,
8 NSW 2052, Australia

9 ² School of BioSciences, University of Melbourne, Parkville, VIC 3010, Australia

10 ³ National Centre for Coasts and Climate (NCCC), School of BioSciences, University of
11 Melbourne, Parkville, VIC 3010, Australia

12 ⁴ Department of Ecology, Environment & Evolution, La Trobe University, Melbourne, VIC
13 3086, Australia

14 ⁵ Department of Behavioural Ecology and Evolutionary Genetics, Max Planck Institute for
15 Ornithology, Seewiesen 82319, Germany

16 ⁶ Department of Earth and Environmental Sciences, Macquarie University, North Ryde, NSW
17 2109, Australia

18 ⁷ Agriculture Victoria Research, Bundoora, VIC 3083, Australia

19 ⁸ Pendoley Environmental Pty Ltd, 12A Pitt Way, Booragoon, WA 6154, Australia

20 ⁹ Australian Institute of Marine Science, Indian Ocean Marine Research Centre, University of
21 Western Australia, Crawley, WA 6009, Australia

22 ¹⁰ Centre for Integrative Ecology, School of Life and Environmental Sciences, Deakin
23 University, Burwood, VIC 3125, Australia

24 ¹¹ College of Science and Engineering, Flinders University, Bedford Park, SA 5042, Australia

25

26

27

28 * corresponding author: m.mayerpinto@unsw.edu.au

29 **ABSTRACT**

30

31 Terrestrial, marine, and freshwater realms are inherently linked through ecological,
32 biogeochemical and/or physical processes. An understanding of these connections is critical
33 to optimise management strategies and ensure the ongoing resilience of ecosystems. Artificial
34 light at night (ALAN) is a global stressor that can profoundly affect a wide range of
35 organisms and habitats and impact multiple realms. Despite this, current management
36 practices for light pollution rarely consider connectivity between realms. Here we discuss the
37 ways in which ALAN can have cross-realm impacts and provide case studies for each
38 example discussed. We identified three main ways in which ALAN can affect two or more
39 realms: 1) impacts on species that have life cycles and/or stages on two or more realms, such
40 as diadromous fish that cross realms during ontogenetic migrations and many terrestrial
41 insects that have juvenile phases of the lifecycle in aquatic realms; 2) impacts on species
42 interactions that occur across realm boundaries, and 3) impacts on transition zones or
43 ecosystems such as mangroves and estuaries. We then propose a framework for cross-realm
44 management of light pollution and discuss current challenges and potential solutions to
45 increase the uptake of a cross-realm approach for ALAN management. We argue that the
46 strengthening and formalisation of professional networks that involve academics, lighting
47 practitioners, environmental managers and regulators that work in multiple realms is essential
48 to provide an integrated approach to light pollution. Networks that have a strong multi-realm
49 and multi-disciplinary focus are important as they enable a holistic understanding of issues
50 related to ALAN.

51

52 **KEY-WORDS:** ALAN, artificial light at night, light pollution, multi-disciplinary, adaptive
53 management, ecological connectivity.

54 INTRODUCTION

55 Artificial light at night (ALAN) is a widespread anthropogenic pollutant that is
56 rapidly increasing in intensity and global distribution. The most current estimates suggest that
57 more than 80% of the human population, and nearly a quarter of the global land area, are
58 exposed to light-polluted skies (Falchi et al. 2016). Consequently, ALAN is reaching most
59 ecosystems globally, with the potential for profound impacts. At its core, ALAN alters
60 natural light-dark cycles, disrupting a key driver of ecological and evolutionary processes
61 (Gaston et al. 2014, Hopkins et al. 2018). Emergent research has linked the presence of
62 ALAN to altered physiology of plants (Bennie et al. 2016) and animals (Dominoni et al.
63 2013) shifts in activity patterns, behaviours, reproduction and survival of animals (Robert et
64 al. 2015, Sanders et al. 2020); disruption of trophic and non-trophic species interactions
65 (Bennie et al. 2015, Gaston et al. 2017); and, significant changes to the structure of ecological
66 communities (Davies et al. 2015, Hölker et al. 2015). The importance and severity of
67 potential impacts of this stressor are increasingly recognised across multiple taxa, habitats
68 and/or ecosystems (Sanders et al. 2020) and there is an increased desire to devise
69 management strategies to minimise ecological impacts of ALAN.

70 A major challenge with mitigating the impacts of ALAN is that, while it is a global
71 environmental pollutant (Falchi et al. 2016) that damages ecological systems (Sanders et al.
72 2020), it is also central to the functioning of modern human society (Edensor 2017).
73 However, beyond natural systems, ALAN can pose public health risks (Pauley 2004) and is
74 energetically and economically costly (Gallaway et al. 2010). Strategies to address the
75 ecological challenges posed by ALAN therefore need to be interdisciplinary, involving
76 researchers (e.g. ecologists, physiologists, social scientists, physicists), managers or
77 regulators (e.g. local councils and government agencies), and practitioners (e.g. urban
78 planners, developers, health specialists, and lighting professionals). While interdisciplinary

79 frameworks have been developed to foster collaboration among researchers, managers and
80 practitioners to better manage urban lighting (e.g. Pérez Vega et al. 2021), they are largely
81 applied within an individual realm (i.e. terrestrial (including land and air), freshwater or
82 marine), rather than considering the potential for light pollution to transcend multiple realms
83 or operate at the realm interface. Current management practices for light pollution do not
84 consider connectivity between realms. Although realms are often considered as separate
85 entities, they are intrinsically linked through ecological, biogeochemical and/or physical
86 processes. Where these linkages are compromised, ecosystem functioning and services might
87 be affected and systems can become less biodiverse and less resilient to change (Beger et al.
88 2010, Field and Parrott 2017). The lack of a multiple-realm integrated approach means
89 outcomes of practices are limited, at best, to small-scale, localised and/or temporary benefits
90 (Threlfall et al. 2021).

91 In this paper, we review examples where ALAN affects two or more realms, directly
92 and/or indirectly. We use the term ‘realm’ as defined by Bugnot et al. (2019), to encompass a
93 group of ecosystems that share common physical and ecological attributes (e.g. the marine
94 realm includes all ecosystems present below the high tide mark while the terrestrial realm
95 includes both air and land). We discuss the consequences of taking a single-realm approach
96 to light pollution management and present a framework to help bridge this gap, incorporating
97 both theoretical and empirical considerations. We also discuss existing challenges and
98 hurdles to studying and managing light pollution. Given ALAN is projected to increase in all
99 three realms in response to continuing human population growth (Kyba et al. 2017), cross-
100 realm management will be critical for ensuring the ongoing resilience of ecosystems
101 (Threlfall et al. 2021).

102

103

104 **Impacts of ALAN on two or more realms**

105 Mitigating the impacts of ALAN and prioritising conservation actions requires
106 consideration of the fundamental interactions among multiple realms (e.g. terrestrial, marine
107 and freshwater) (Beger et al. 2010). Realms may be linked through ecological,
108 biogeochemical, or physical processes (or combinations of these), including the movement of
109 organisms, materials and energy between ecosystems and realms; this link is broadly referred
110 to as ecological connectivity (Taylor et al. 1993). Shifts in ecological connectivity through
111 the disruption of daily, seasonal or cyclic movement of organisms or resources will likely
112 have consequences across multiple realms. For example, changes to predation and foraging
113 behaviours at the level of the individual or community (e.g. species diversity and richness)
114 can have cross-realm implications due to trophic cascades and linked changes in ecosystem
115 function through nutrient cycling or pollination. This is particularly true if the organisms
116 involved typically function across realm boundaries. Similarly, individual-level shifts can
117 have cross-realm ecological consequences if the species in question has life histories or
118 migratory patterns that traverse multiple realms, such as the two case studies we discuss
119 below, salmon (freshwater juveniles, marine adults) and secondarily aquatic insects (aquatic
120 juveniles, terrestrial adults).

121 ALAN-driven impacts include changes in the phenology, growth form and resource
122 allocation of plants (Bennie et al. 2016), as well as the behaviour, physiology, distribution
123 and survival of animals (Brüning et al. 2011, Perkin et al. 2014, Bolton et al. 2017, Fobert et
124 al. 2019, Willmott et al. 2019, Aulsebrook et al. 2020). Mechanisms driving such impacts,
125 which could then directly or indirectly affect other realms, include changes in the flux of
126 inorganic and organic material. Changes in oxygen and nutrient fluxes, for example, can
127 potentially directly impact land, sea and freshwater habitats (Hölker et al. 2015, Grubisic et
128 al. 2017). Indirect effects can be driven by bottom-up or top-down processes. Bottom-up

129 processes occur when effects on primary producers (e.g. algae or autotrophic microbes) affect
130 populations at higher trophic levels through changes in resource availability. For example, an
131 increase or reduction in the diversity and abundance of aquatic insects due to ALAN is
132 expected to have implications for terrestrial consumers that rely on aquatic prey, such as
133 spiders, birds and bats (Baxter et al. 2005, Zapata et al. 2019). Alternatively, changes may be
134 driven by top-down processes, arising from impacts on, for example, the survival or
135 behaviour of herbivores and/or predators. Consequences of such changes are varied and
136 dependent on the magnitude of change, but may result in loss of biodiversity (Bowyer et al.
137 2005).

138 ALAN is also likely to have cross-realms consequences if the effects occur within
139 ecosystems that link multiple realms – i.e. transitional zones – such as estuaries and coastal
140 wetlands, which are at the intersection of freshwater, marine, and terrestrial realms. Such
141 zones, and the organisms that inhabit them, tend to be disproportionately affected by ALAN,
142 because urban settlements where ALAN is prevalent, are often developed near waterways
143 (Kummu et al. 2011).

144 Rapid changes in the environment, such as those caused by ALAN, can alter
145 environmental cues used by many animals to select optimal habitats that maximise their
146 fitness (Hale and Swearer 2016, Swearer et al. 2021). Such ‘ecological traps’ can promote
147 disruptions or alterations in the movement patterns of organisms, resulting in increased risk
148 of mortality and/or shifts in trophic interactions (Schlaepfer et al. 2002), with potential
149 implications for multiple realms. Examples that have major consequences for species in
150 multiple realms include turtle hatchlings crawling inland towards artificially-lit beach fronts
151 instead of heading seaward (Witherington and Bjorndal 1991), or once reaching the ocean,
152 swimming towards lights on water such as on boats, piers or other nearshore infrastructure
153 (Thums et al. 2016, Wilson et al. 2018). This can have serious implications for the health of

154 seagrass meadows, with likely flow-on effects on the diversity supported by these important
155 habitats (e.g. Hernández and van Tussenbroek 2014). Similarly, terrestrial insects can fall for
156 these ecological traps if they attempt to land or lay eggs on impervious concrete surfaces that
157 reflect light and thus are mistaken for water (Horváth et al. 2009). This failure to lay eggs in
158 the appropriate habitat can impact offspring survival with cross-realm impacts. While
159 ecological traps do not inherently have cross-realm impacts, the impacts of ecological traps
160 created by ALAN can cross realm boundaries if it disrupts species interactions or movements
161 that occur across more than one realm (see Box 1).

162 Based on the above, we have identified three broad ‘ways’ where there is evidence, or
163 likelihood, of ALAN- related cross-realm impacts; they operate by impacting:

164 1) species that move across realms, through life cycles and/or stages or migratory patterns
165 that occur in two or more realms, such as diadromous fish and many insects, as well as
166 marine reptiles, mammals (e.g. seals) and birds (e.g. penguins and albatross) that are tied to
167 land for breeding and/or resting; 2) species interactions, such as predator-prey interactions,
168 that occur across realm boundaries; and 3) transitioning zones or ecosystems such as coastal
169 wetlands and estuaries, which inherently link realms. The impacts of ALAN on these cross-
170 realm linkages can be further altered or exacerbated if light pollution is acting as an
171 ecological trap (see Box 1). Below, we provide examples or case studies, where possible,
172 stating observed or inferred/likely effects of ALAN, and discuss their cross-realm
173 consequences.

174

175 ***1) Impacts on species with life cycles/stages across two or more realms***

176 The life cycles of many organisms occur in two or more realms. Examples include
177 animals whose juveniles are aquatic while adults are predominantly marine or terrestrial, or
178 marine animals that breed on land or in freshwater systems. Impacts of ALAN on any one

179 stage are, therefore, predicted to have carry-over effects on subsequent life-stages,
180 consequently impacting different realms. We use two case studies to illustrate this, one on
181 salmon (Salmonidae) and the other gives a broader overview of secondarily aquatic insects,
182 such as dragonflies and mayflies.

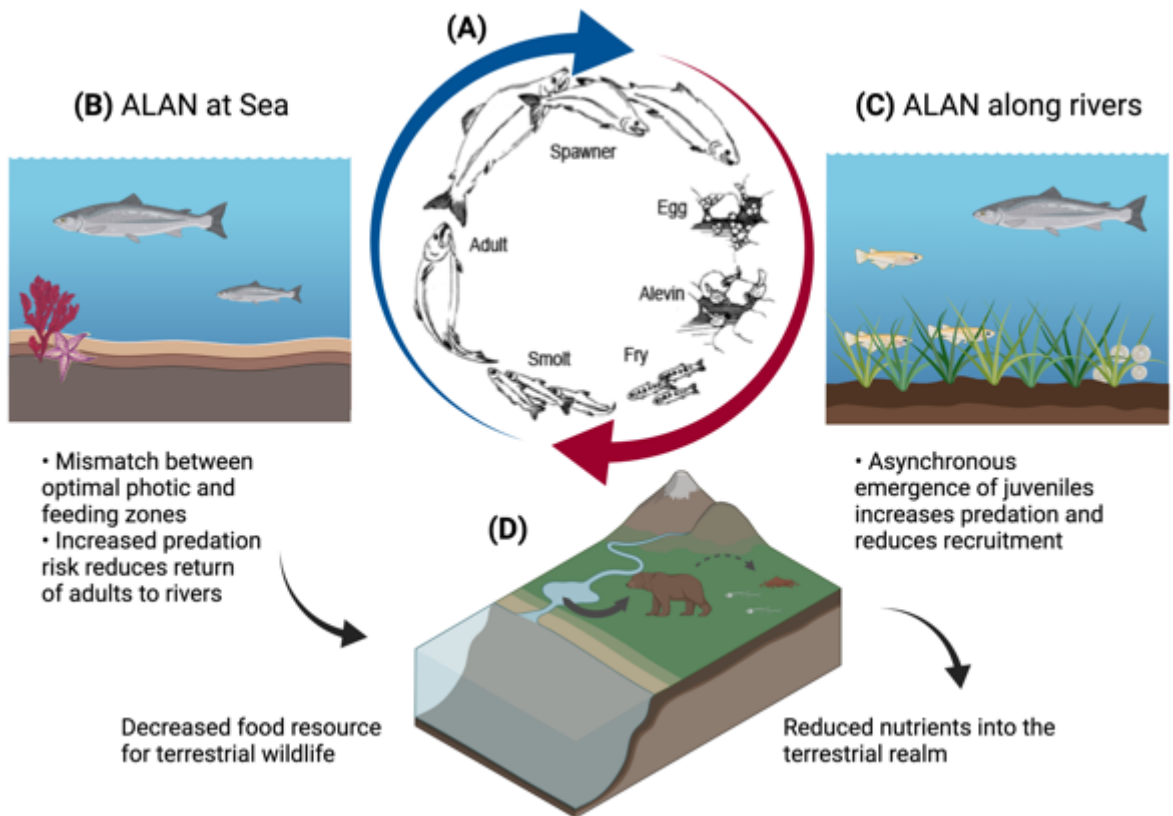
183

184 *Case study 1 – Salmon, a vector of energy and nutrients across realms*

185 Salmon, including the Atlantic (*Salmo salar*) and Pacific Salmon (*Oncorhynchus* spp.),
186 are anadromous fish, meaning they spend their juvenile phase (e.g. alevins, fry, and parr) in
187 rivers, before migrating to the ocean as smolts (1-3 yr old juveniles that are physiologically
188 adapted for sea water) to feed, grow, and mature. Adults then return to freshwater systems for
189 spawning (Figure 1). ALAN has demonstrable impacts on several life-stages of salmon
190 including fry (Riley et al. 2013, Riley et al. 2015) and smolts (Riley et al. 2012). For
191 example, emergence of juvenile Atlantic salmon in streams is usually mediated by
192 environmental cues, such as presence of predators (Jones et al. 2003). Fry are highly
193 vulnerable to predation, and synchronous emergence can increase their chances of survival
194 (Brännäs 1995). ALAN associated with human populations along river systems is linked to
195 asynchronous nocturnal emergence, disrupted dispersal and decreased weight of fry in the
196 freshwater realm (Riley et al. 2013). This has knock-on effects for population recruitment and
197 predation risk given that both synchronous nocturnal emergence and dispersal are posited as
198 predator avoidance mechanisms (Riley et al. 2013). At the smolt stage in the marine realm,
199 ALAN associated with aquaculture practices, for example, can alter the vertical movement of
200 salmon, causing trade-offs between preferred light and temperature levels, feeding, and risk
201 perception (Oppedal et al. 2011). Furthermore, a field experiment showed that smolt
202 populations exposed to ALAN from streetlights presented altered migratory behaviour,

203 potentially impacting their fitness and/or predation risk (Riley et al. 2012), with likely
204 consequences for the total biomass of fish surviving to the ocean life-stage.

205 In addition to the direct impacts on salmon, these fish are important vectors in
206 transporting energy and nutrients between the ocean, freshwater and terrestrial environments
207 (Gende et al. 2002); therefore, impacts in one environment will likely have cross-realm
208 consequences (Figure 1). For example, migrating adult salmon serve as a food resource for
209 terrestrial wildlife as they travel upstream to spawn. Bears alone move up to 90% of all
210 salmon biomass to land, sometimes hundreds of meters from their stream of origin
211 (Reimchen 2000). Salmon-derived minerals and nutrients are further spread in the terrestrial
212 environment through bear urine and faeces as these mammals move throughout the riparian
213 and upland forests (Hilderbrand et al. 1999). Salmon also support freshwater systems by
214 providing nutrients from their carcasses following spawning (Juday et al. 1932) and play an
215 important role in the marine food-web during their migratory stage to the sea (Gende et al.
216 2002). Therefore, efforts to mitigate the impacts of ALAN on salmon, that are solely focused
217 in one realm may be ineffective and economically wasteful if impacts from/in other realms
218 are not considered.



219

220 Figure 1 - Schematic figure showing the potential cross-realms impacts of ALAN due to

221 effects on different life stages in salmon species. (A) Salmon spend their juvenile phase in

222 rivers before migrating to sea to grow and mature. To complete their life cycle they must

223 return to the river to spawn. (B) ALAN at sea alters vertical movement of fish resulting in a

224 mismatch between preferred light levels and optimal feeding zones. Additionally, ALAN

225 results in increased predation of fish at sea and hence a decrease in adults returning to rivers.

226 (C) ALAN along rivers disrupts synchronous emergence of juveniles resulting in increased

227 predation which then reduces the recruitment of smolts out to sea. This reduction in adults

228 returning to rivers and smolts migrating to sea results in trophic effects in both realms. (D)

229 Illustrates one trophic effect in the terrestrial environment with reduced food resources for

230 bears resulting in reduced nutrients into the terrestrial environment. Image created with

231 BioRender.com.

232

233 *Case study 2 - Aquatic insects (with terrestrial adults)*

234 Secondarily aquatic insects - those with an aquatic egg and juvenile phase and a
235 terrestrial adult phase - are proposed as ideal bioindicators to assess the impact of cross-realm
236 (aquatic and terrestrial) environmental change due to their sensitivity to anthropogenic
237 stressors (Villalobos-Jimenez et al. 2016). However, there is an overall lack of direct
238 evidence for how impacts in any one of these realms can influence others. Moreover, there is
239 surprisingly little information regarding the specific impact of ALAN on the independent life
240 history stages of secondarily aquatic insects: in the largest review of urban impacts on
241 dragonflies, ALAN was not even included (Villalobos-Jimenez et al. 2016). Nevertheless, the
242 overall life-history knowledge we have on these organisms, coupled with the current existing
243 information on ALAN impacts on insects and their habitats more broadly, allows us to
244 infer/hypothesise likely cross-realms impacts. For instance, the effect of variation in
245 moonlight on adult insect activity has been long documented (Williams and Singh 1951) and
246 it is well recognised that artificial lighting is attractive to adult insects, with this behaviour
247 being commonly exploited when trapping potential pests (Shimoda and Honda 2013).
248 Furthermore, increasing evidence suggests that artificial light at night will have multiple
249 negative consequences for stream and riparian habitats (Perkin et al. 2011). Therefore, we
250 discuss ways in which ALAN is likely to have cross-realms impacts through effects on both
251 the terrestrial and aquatic life stages of secondarily aquatic insects.

252 Ecologically, dragonflies, mayflies and mosquitoes are classic examples of
253 secondarily aquatic insects that have a relatively short terrestrial adult phase and a protracted
254 aquatic egg and larval phase. The transition from the juvenile aquatic environment to the
255 terrestrial adult environment is varied and taxon-specific. For example, prior to their final
256 moult, dragonfly nymphs typically move up out of the water (usually at night) onto a branch
257 or other structure where they eclose and emerge as air-breathing terrestrial adults. Mosquitoes

258 remain in the aquatic environment emerging directly into the terrestrial environment as
259 adults, typically remaining at the surface to allow their wings to dry and harden. Mayflies are
260 hemi-metabolous and thus do not have a pupal stage, instead they either emerge into the
261 terrestrial environment as a winged subadult (or sub-imago) and then rapidly moult to adults.

262 The mechanisms that promote ALAN-specific cross-realm impacts for secondarily
263 aquatic insects are varied. Point sources of ALAN close to streams or water bodies may
264 change patterns of dispersal (geographic or temporal; Manfrin et al. 2017) and/or act as
265 ecological traps for newly emerging adults (Eisenbeis et al. 2006, Perkin et al. 2011). Such
266 behaviours may lead individuals away from the aquatic environment required for mating and
267 egg laying (Eisenbeis et al. 2006, Perkin et al. 2011) and into an environment where the risk
268 of predation is increased (Davies et al. 2012). Some species (e.g. dragonflies) are also
269 positively polarotactic, using horizontally polarized light to locate suitable water bodies for
270 mating and egg laying (Kriska et al. 2009). In areas with anthropogenic sources of polarised
271 light (asphalt surfaces, vertical glass and even vehicles) these behaviours can be disrupted
272 leading surviving adults to aggregate and females to oviposit on suboptimal non-aquatic
273 surfaces where juvenile survival is reduced or non-existent (Horváth et al. 2014). Similar
274 effects are documented for mayflies and caddisflies, whose attraction to anthropogenic
275 sources of polarised light at night can reduce reproductive success and increase risk of
276 predation by light-attracted insectivores, such as birds, lizards or spiders (Robertson et al.
277 2010, Szaz et al. 2015).

278 Assuming eggs are laid in a body of water, the protracted aquatic juvenile phase may
279 be vulnerable to the impact of ALAN. Evidence from other insects suggests aquatic juveniles
280 may be directly attracted to external light sources, leading to shifts in foraging or other
281 activity patterns (Kühne et al. 2021), which may result in increased predation risk (Manfrin et
282 al. 2018). Prolonged juvenile exposure to ALAN may also negatively impact growth,

283 development and survival, as shown in crickets (*Teleogryllus commodus*) (Durrant et al.
284 2018) or reduce fecundity (McLay et al. 2017, Willmott et al. 2018). Finally, ALAN may
285 have indirect impacts by promoting shifts in the aquatic community structure, reducing
286 availability of prey (Hölker et al. 2015). Ultimately, the degree to which exposure to ALAN
287 results in selection of particular juvenile phenotypes that survive to the adult stage (Hopkins
288 et al. 2018). The impact this has on juvenile or adult life history decisions, including flight to
289 light behaviour is unknown. Nevertheless, impacts are expected. Hence, such knowledge is
290 critical if we are to understand the fitness consequences for species, such as secondarily
291 aquatic insects, that cross multiple realms and their knock-on effects.

292

293 **2) *Impacts on species interactions that involve two or more realms***

294 To date, most studies on ALAN focus on the evolutionary and ecological
295 consequences of ALAN at the level of the individual or population within a single realm (e.g.
296 terrestrial, Sanders et al. 2020). However, the loss of, or changes in, species within a system
297 can affect an entire cross-realm network, through altered competition and/or food-web
298 interactions, with unpredictable consequences for communities, ecosystems (Eklöf and
299 Ebenman 2006) and other, connected, realms (Bugnot et al. 2019). Below, we highlight two
300 case studies where observed or inferred effects of ALAN for one species or group are
301 expected to affect multiple realms through species interactions and knock-on effects.

302

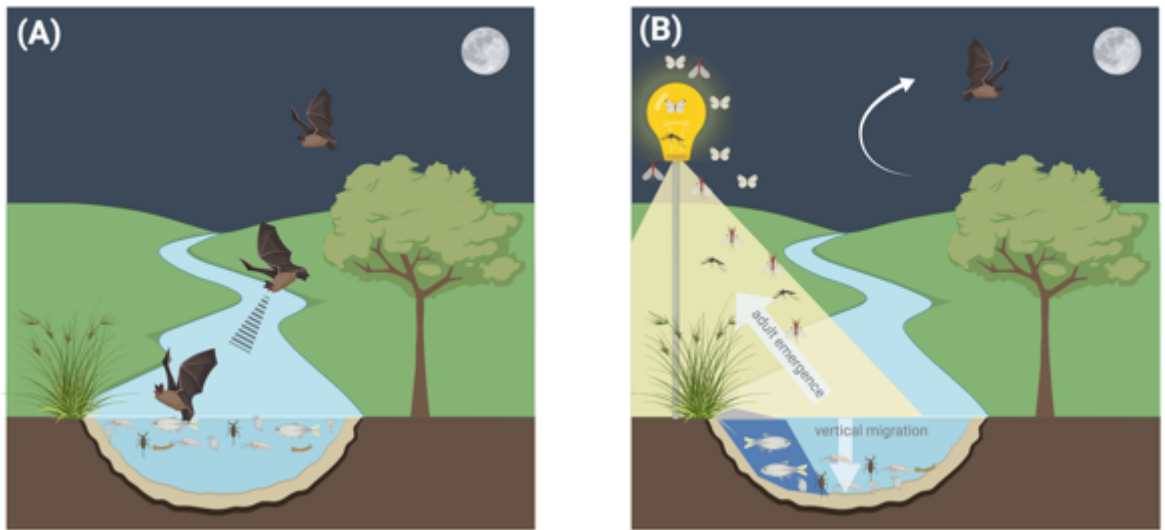
303 *Case study 3 - Fishing bats: terrestrial mammals specialised for feeding in aquatic*

304 *ecosystems*

305 Worldwide, there are 16 species of fishing or trawling bats (e.g. from the genus
306 *Myotis*). This group has ecological and foraging specialisations that make them reliant on
307 both terrestrial and aquatic realms (Campbell 2011). Fishing bats roost in caves, aqueducts,

308 bridges, tunnels and tree cavities in the vicinity of water sources (Campbell 2009, Gorecki et
309 al. 2020) and forage exclusively on aquatic prey using their feet to trawl the surface of water
310 for fish and aquatic insects (Dwyer 1970, Law and Urquhart 2000, Campbell 2007). An
311 emerging issue facing fishing bats that depend on surface foraging in waterways is the shifts
312 in the aquatic prey behaviour driven by ALAN. Such shifts often result in changes in the
313 distribution and behaviour of prey, which affect the bats' ability to forage (Figure 2). Fishing
314 and trawling bats cannot detect submerged prey (Suthers 1965) and instead rely on
315 echolocation of water surface irregularities created by fish and aquatic invertebrates
316 (Thompson and Fenton 1982). Light is a critical cue for diel vertical migration: during the
317 day, aquatic invertebrates (potential prey items) move downwards from the water's surface to
318 deeper water, while during the night, prey move upwards to the surface (Perkin et al. 2011,
319 Mehner 2012). Darkness is also a cue for the emergence of adult aerial invertebrates from the
320 aquatic realm (Manfrin et al. 2017). Under ALAN, nocturnal vertical migration of
321 invertebrates to the surface is reduced and fishing bats are limited in their ability to detect
322 their aquatic prey. This reduction in foraging is compounded by the absence of smaller fish
323 which in the presence of ALAN hide in the shadows.

324 To understand the impacts of ALAN on waterways, consideration of the direct
325 impacts of ALAN on movement patterns of resident aquatic species, as well as the direct and
326 indirect impacts on terrestrial species that feed on those aquatic species is required. For
327 example, fishing bats appear to be light averse and actively avoid lit areas, possibly due to
328 increased risk of predation (Straka et al. 2016), which reduces their ability to capitalise on the
329 increased emergence of some aquatic insects and the attraction of aerial invertebrates to
330 lights. Accordingly, experimental research highlights a reduction in recorded feeding
331 attempts when waterways are lit (Haddock 2019), suggesting they are unable to switch prey
332 resources to take advantage of the abundance of aerial insects attracted to the light source.



333

334 Figure 2 - (A) Schematic figure depicting the aquatic ecosystem with fishing bats under
335 natural light (B) and how artificial light at night influences prey species. As artificial light is
336 introduced, aquatic prey species migrate into shadows, sediment or to greater depths, making
337 them unavailable to bats. Additionally, some aquatic insects emerge as aerial adult forms that
338 are attracted to light. Fishing bats avoid lit areas and cannot switch foraging strategies to take
339 advantage of the new aerial prey that is attracted to lights. Image created with
340 BioRender.com.

341

342 *Case study 4 – Shifting energy flows between realms via impacts on orb-web spiders and*
343 *aquatic insect communities*

344 Resource exchange from terrestrial to aquatic realms is an intrinsic facet of riparian
345 habitats (Baxter et al. 2005). Spiders are important predators in riparian zones and can obtain
346 more than 50% of their nutrition from aquatic sources, especially insects (Collier et al. 2002).
347 Therefore, effects of ALAN on the diversity, abundance and distribution of spiders (both
348 free-living and web-building), and/or the community of aquatic insects in riparian zones can
349 alter cross-realm fluxes, with important regional and global implications for both terrestrial

350 and aquatic realms (Manfrin et al. 2017). The consequences of these effects of ALAN depend
351 on the time-scale considered and may be sex-specific. For example, short-term (two-month)
352 exposure to ALAN increased the abundance and body mass of riparian long-jawed orb
353 weavers (family Tetragnathidae) (Parkinson et al. 2020). These effects were more
354 pronounced for females compared to males and were concordant with greater numbers of
355 prey items captured in spider webs under ALAN compared to webs under natural night-time
356 conditions. However, a comparable but longer-term study (one year) found that although
357 spider density initially increased (as in the previous study), there was a long-term decrease in
358 spider density, as well as a decrease in the emergence of aquatic insects (Meyer et al. 2013).
359 Together, these studies suggest that aggregation of predators and prey around ALAN can
360 increase predation on emerging aquatic insects and so reduce the transfer of biomass from
361 aquatic to terrestrial systems through riparian zones. Concurrently, this would shift biomass
362 from dark areas into artificially illuminated areas and dramatically shift the distribution,
363 overall abundance, and diversity of insect communities reducing their abundance as prey
364 (Perkin et al. 2014, Manfrin et al. 2017, Parkinson et al. 2020). Therefore, by altering both
365 the abundance and predation success of terrestrial predators, as well as the distribution and
366 abundance of aquatic prey, ALAN can drive shifts in predator-prey interactions across realm
367 boundaries, altering flows of energy between aquatic and terrestrial systems, with important
368 consequences for both realms.

369 ***3) Impacts on transition zones***

370 Estuaries and coastal wetlands are critical transition zones that link freshwater
371 habitats with marine and terrestrial environments (Levin et al. 2001). These zones perform
372 important ecological functions such as nutrient cycling and regulation of water and nutrient
373 fluxes between realms (Levin et al. 2001). Riparian zones are also at the interface of
374 terrestrial and aquatic systems and support high biodiversity, as well as key ecosystem

375 functions through biogeochemical cycling (Naiman and Decamps 1997). Therefore, impacts
376 of light pollution on these critical transition zones are likely to cross ecosystem boundaries,
377 affecting two or more realms, with multiple consequences for multiple functions of
378 ecosystems and the services they underpin.

379 Natural light at the air-water interface is a key factor linking terrestrial and aquatic
380 realms. The amount of light that reaches the water surface in freshwater or coastal systems,
381 depends on the surrounding terrestrial habitat: structurally complex terrestrial environments,
382 such as forested riparian zones, reduce the amount and colour of light reaching the water
383 surface (Endler 1993). Organisms also vary extensively in their sensitivities to multiple light
384 properties (Gaston et al. 2012, Land and Nilsson 2012), and transition zones support several
385 specialised species that have adapted to these complex lighting environments. For example,
386 some estuarine fish species that can live in highly turbid waters with low ambient light levels
387 due to high loads of suspended material, such as the flathead grey mullet (*Mugil cephalus*),
388 have morphological traits that support dim-light (i.e. scotopic) vision, such as high rod
389 density in the retina (Zapata et al. 2019). The freshwater three-spine stickleback
390 (*Gasterosteus aculeatus*) also has highly specialised visual sensitivity important for mate
391 selection in clear versus tannin-stained lakes (Boughman 2001). Due to their evolutionary
392 history, organisms inhabiting transition zones may be more sensitive to the presence of
393 ALAN that modifies the unique light environment in which they have evolved (Sullivan et al.
394 2019). A further problem is that transition zones tend to be disproportionately affected by
395 ALAN, since many urban settings, where ALAN is prevalent, are developed near waterways
396 (Kummu et al. 2011). Transition zones, therefore, are significant sites for understanding and
397 managing cross-realm impacts of ALAN, both due to the vulnerability of organisms
398 inhabiting these zones, and the prevalence of light pollution near waterways.

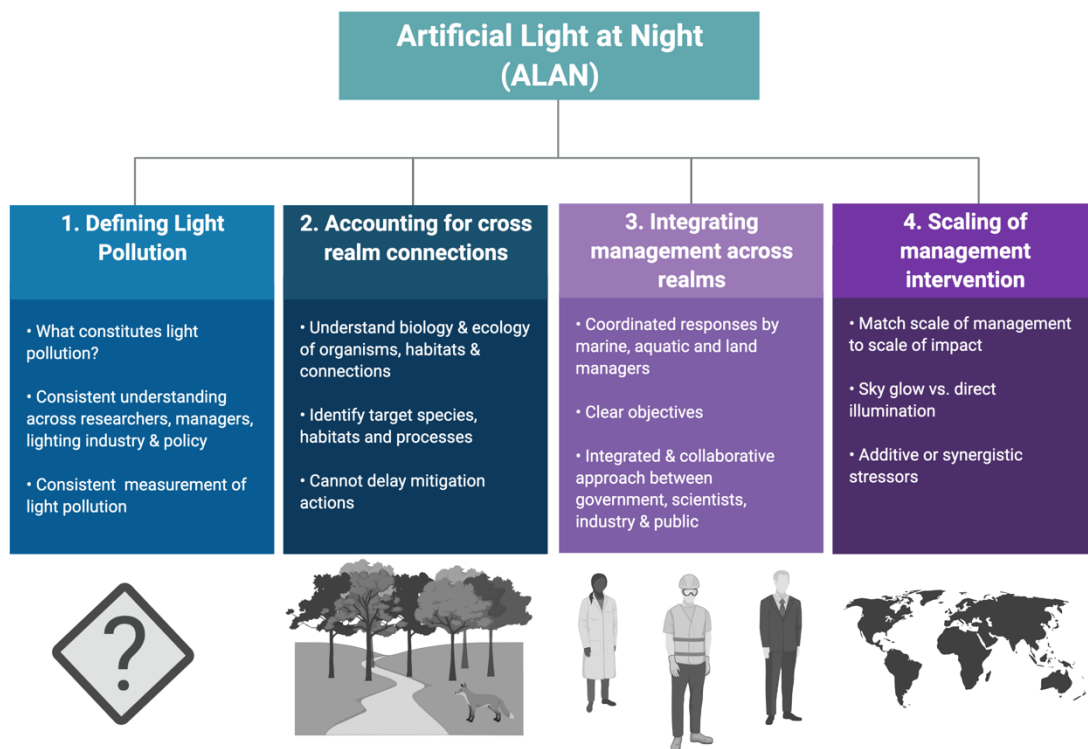
399 The orb web spiders and aquatic insects example outlined above (Case study 4)
400 illustrates how shifts in the flow of resources in riparian zones – the interface between land
401 and rivers or streams – can have impacts across multiple realms. Additionally, in their recent
402 comprehensive review, Zapata et al. (2019) outlined a multitude of ways ALAN can affect
403 estuaries and highlighted potential cross-realm implications. For example, ALAN-induced
404 delays in the leaf fall of deciduous trees (Bennie et al. 2016) can in turn reduce the input of
405 nutrients from leaf detritus into aquatic systems, causing potential shifts in the
406 biogeochemistry of aquatic systems (Zapata et al. 2019). Sullivan et al. (2019) also recently
407 demonstrated the impacts of ALAN on riparian systems through shifts in the community
408 structure of invertebrates, consequently altering the flows of energy between aquatic and
409 terrestrial systems. Given these direct examples and published review of the impacts of
410 ALAN on transition zones and flow-on effects across realms, we have not provided case
411 studies here to further illustrate this mechanism. Instead, we want to highlight the importance
412 of prioritising transition zones for management actions to limit the impacts of light pollution
413 across multiple realms.

414

415 **CHALLENGES AND PRACTICAL SOLUTIONS FOR RESEARCH AND** 416 **MANAGEMENT OF ALAN**

417 Several challenges exist that need to be addressed for the impacts of light pollution to
418 be effectively understood and managed, both within and across realms. A major difficulty
419 (and potential point of contention) encountered when dealing with cross-realm issues is
420 determining the boundaries for management and governance (Pittman and Armitage 2016).
421 For example, land-based sources of ALAN may indirectly influence the productivity of
422 aquatic systems through its impact on nutrient inputs from terrestrial sources through e.g.
423 changes in the leaf fall patterns of deciduous trees. In this case, areas are separated by

424 physical and jurisdictional boundaries (e.g. land and coastal managers) and potentially social
 425 boundaries (different communities or social networks). Here, we propose a framework for
 426 cross-realm management, which builds on previous frameworks for conservation and
 427 management across-realms (e.g. Beger et al. 2010, Alvarez-Romero et al. 2015, Giakoumi et
 428 al. 2019, Threlfall et al. 2021), but with a specific focus on light pollution (Figure 3).



429
 430 Figure 3 - Proposed framework to integrate cross-realms considerations into the study and
 431 management of light pollution. Image created with BioRender.com.

432
 433 ***Challenges and practical solutions***

434 ***1 - Defining light pollution***

435 One of the main challenges for driving practical solutions to manage ALAN is
 436 agreeing to a collective understanding of how and when lighting should be defined as
 437 pollution (Schulte-Römer et al. 2019). Here, we define light pollution as light introduced into

438 the environment by humans at intensities that are higher than the natural level at that time for
439 the given environment and that has the potential to cause harm to humans and/or the
440 environment. In a recent analysis, Schulte-Römer et al. (2019) found that light pollution
441 experts (including scientists and managers) had a stronger and more consistent view of what
442 constitutes light pollution than lighting professionals (such as lighting designers, urban
443 planners and engineers). Importantly, however, both groups had very skewed views when
444 considering potential issues caused by light in areas where it is 'unwanted', depending on the
445 habitat or realm. Approximately 90% of light pollution experts (n = 89 respondents)
446 considered light to be pollution when it obscures the visibility of stars, or when fixtures were
447 installed close to observatories. In contrast, only 66% of those surveyed considered lighting
448 as pollution when it was installed close to bodies of water, and, among lighting professionals,
449 this dropped to only 17% (n = 67 respondents). These results highlight a common
450 misconception, and a massive global problem, namely, that light is a 'land' problem rather
451 than of fundamental significance for all ecosystems on earth. These findings also ignore the
452 critical need for fluctuating light levels (both day and night) that have characterised the
453 evolutionary history of that life. Therefore, the first steps to successfully managing light
454 pollution within and across realms are to (i) raise awareness of the importance of fluctuating
455 light regimes for ecological process; (ii) enhance understanding of the impacts of artificial
456 light across all realms: terrestrial, freshwater and marine environments; (iii) broaden
457 knowledge regarding the impact that light within one realm can have for biodiversity and
458 ecosystem function within other realms and (iv) understand the 'acceptable' levels of ALAN
459 for both the local ecological communities and society (i.e. trade-offs between ecological
460 impacts and societal needs or desires). Critically, this needs to include multiple stakeholders,
461 including the general public.

462

463 *2 - Accounting for cross-realm connections*

464 The next step in managing light pollution across realms is to understand the biology
465 and ecology of organisms and habitats of interest and their potential linkages, so that
466 management interventions can more fully account for connections across realms. Ideally, the
467 extent of the impact of ALAN on target individuals, populations, habitats and systems, as
468 well as the mechanisms driving these changes, will be well-known within and across realms.
469 However, we acknowledge that, unfortunately, the current state of habitat degradation
470 worldwide and rapid expansion of ALAN means that we cannot afford delaying mitigation
471 actions until the impacts, or even the potential unintended risks of management interventions,
472 are fully understood (Mayer-Pinto et al. 2019). Therefore, we need to keep gathering the -
473 still much needed - scientific information on the effects of ALAN, within and across realms,
474 while, at the same time, implementing local, regional and global best practice guidelines to
475 prevent or lessen such impacts.

476

477 *3 - Integrating management across realms*

478 Another key issue with cross-realm management of light pollution is the lack of
479 collaboration between different stakeholders and the existence of methodological disparities
480 across realms. To address this, it is important to clearly specify the desired outcomes and to
481 standardise approaches/methodologies regarding ALAN and its impacts across realms. The
482 compartmentalisation that can exist within governance structures, such as within and between
483 local, state/territory and federal government agencies often results in a lack of consistency in
484 management decisions across realms. This can be due to poor communication, differing and
485 potentially competing priorities and a lack of collaboration among the sectors and agencies
486 responsible for planning and environmental protection in the different realms; a lack of
487 spatial data on cross-realm processes; and difficulties arising in adapting existing decision-

488 tools and coordinating different governance systems (Alvarez-Romero et al. 2015 and
489 references therein).

490 There are some key general steps, as outlined by Alvarez-Romero et al. (2015),
491 Bugnot et al. (2019) and Threlfall et al. (2021), among others, to successfully implement
492 cross-realm management strategies. First and foremost, a clear objective is necessary, i.e.
493 what are the desired outcomes? For issues pertaining to light pollution, these can include
494 minimising the effects of ALAN on ecologically, culturally and/or commercially important
495 target species/groups or a target area (e.g. a transition zone, migratory pathways or a
496 protected area). This requires an integrated and collaborative approach with policy makers,
497 regulators, scientists, lighting designers, developers and the general community, including
498 First Nations People, so that potential conflicting interests are identified, and solutions are
499 devised accordingly. Consequently, we need to not only unify terminologies and agree on
500 desired outcomes (Webb 2012, Bugnot et al. 2019), but ideally, understand potential
501 thresholds of ‘acceptable’ artificial light levels across different species and realms.

502 Determining ALAN thresholds, however, requires standardised measurements of light
503 *per se*. Currently, there is great inconsistency in instrumentation and light parameters within
504 and across realms. Discrepancies in lighting measurements exist for valid and practical
505 reasons – e.g. the measurement and instrument used needs to match the scale of both the light
506 pollution being measured (i.e. direct source vs skyglow) and the ecological or biological
507 response of interest (e.g. insect attraction to a street light vs bird migration). Moreover, as far
508 as we know, there is not yet available affordable and easy-to-use instrumentation to
509 adequately measure light levels under water. However, there is a clear and urgent need to
510 standardise, where possible, the measurement of light pollution, so that outcomes are
511 comparable and applicable across realms (see Box 3 for further discussion). It is important to
512 note, however, that knowing relevant light ‘levels’ is not enough for effective management

513 for ecological outcomes. At the extreme, any light that is not natural in its origin is likely to
514 interfere with ecological process. Thus, perhaps of greater importance, we need to be able to
515 measure and understand how light properties (including spectra and intensity) affect
516 organisms and habitats in multiple realms. Standardising how and which properties of light
517 are measured will facilitate communication of clear and specific recommendations (including
518 biologically relevant thresholds) between researchers, practitioners and managers. This will
519 permit informed decision making when taking into account potential impacts across different
520 habitats and realms and the risks we are willing to take when night-time illumination is
521 unavoidable and/or socially desirable.

522

523 *4 - Scaling of management intervention*

524 Ultimately, there is a need to match the scale of the management intervention to the
525 scale of impact (Threlfall et al. 2021). Light pollution impacts occur at the landscape scale,
526 and include impacts caused by sky glow, light scattered in the atmosphere (Cinzano et al.
527 2001, Falchi et al. 2016), and those caused by direct illuminance from light sources (e.g.
528 streetlights). Impacts caused by direct illuminance are, in theory, easier to mitigate, than
529 impacts caused by sky glow – which can be an issue even tens (and possibly hundreds) of
530 kilometres from urban light sources (Gaston et al. 2012) and require management
531 interventions at much larger, landscape level, scales to prevent or mitigate cross-realm
532 impacts. For example, research has shown that light pollution can spill into otherwise
533 protected areas up to 15 km from urban centres (McNaughton et al. 2021). Additionally, a
534 recent study has highlighted the potential for synergistic interactions between sky glow and
535 direct illuminance (Dickerson et al, unpublished data). Management actions therefore need to
536 consider, whenever possible, multiple spatial scales to mitigate light pollution and avoid

537 cross-realm impacts. Extensive examples on specific interventions and management
538 strategies can be found in the literature (Gaston et al. 2012, DAWE 2020).

539 Light pollution is just one of a multitude of anthropogenic stressors associated with
540 urbanisation (Dominoni et al. 2020), which can also cross realm boundaries. Therefore,
541 management interventions should also consider potential additive or interacting impacts from
542 multiple stressors (Hale et al. 2017). For example, ALAN and night-time warming have non-
543 additive interactive effects on the predation of aphids by lady beetles, decreasing aphid
544 population densities (Miller et al. 2017). Similarly, particular traits in birds can be impacted
545 by both ALAN and noise pollution: light pollution is associated with advancement in
546 reproductive phenology of several species of birds while noise decreased clutch size of
547 closed-habitats (i.e. forests) birds (Senzaki et al. 2020). Interactive effects of stressors remain,
548 however, poorly understood. Understanding, or at a minimum identifying, other stressors that
549 may interact with or act simultaneously with ALAN will enhance cross-realm management
550 outcomes. Moreover, climate change adds additional challenges to cross-realms studies since
551 it increasingly modifies key land-sea ecological and social processes, therefore increasing the
552 urgency for transboundary management initiatives.

553

554 **CROSS-REALMS MANAGEMENT SUCCESS**

555 There have been few examples of successful of management of ALAN which have
556 resulted in a reduction of cross-realm impacts, and most of these examples involved
557 management interventions that targeted a single species. Successful examples include the
558 mitigation of impacts on shearwaters at Phillip Island, Melbourne (Rodríguez et al. 2014,
559 Rodríguez et al. 2017) and on nesting marine turtles (discussed in more detail here). Marine
560 turtles have complex life histories that cross marine and terrestrial realms, and are considered
561 key indicators of ecosystem health (Haywood et al. 2019). In Australia, marine turtles are

562 protected under environmental legislation. As light pollution can reduce the reproductive
563 viability of turtle stocks by disrupting critical behaviour such as the ability of hatchling
564 marine turtles to successfully reach the ocean (Witherington and Bjorndal 1991), all actions
565 in Australia that involve artificial light that is likely to impact marine turtles must be referred
566 for environmental assessment. Proponents must demonstrate, via formal risk assessments,
567 how the impact of ALAN on all age classes of marine turtles will be mitigated and adaptively
568 managed. Light in nearshore waters (e.g. boats on anchor, jetties, coastal lighting, etc), for
569 instance, influence the offshore dispersal of hatchlings in the critical minutes and hours after
570 they leave the beach. Attraction to artificial lights increases the time hatchlings spend
571 crossing predator rich nearshore waters before reaching the safety of deep water offshore,
572 thus increasing their vulnerability to predation (Harewood and Horrocks 2008, Thums et al.
573 2016, Wilson et al. 2018); and as predators are also attracted to the same lights, predation
574 pressure can be high. Mitigation measures that benefit marine turtles have been summarised
575 in the National Light Pollution Guidelines for Wildlife Including Marine Turtles, Seabirds
576 and Migratory Shorebirds (DAWE 2020) and include management of the physical aspects of
577 the light, such as intensity (lumen output), colour (wavelength) and elevation above dark
578 horizons behind the beach, as well as the maintenance of dark zones between turtle nesting
579 beaches and light sources, and shielding and targeting of light fixtures to avoid direct
580 visibility and limiting sky glow (DAWE 2020). Given light pollution sources that can affect
581 turtles can be both marine and terrestrial, management actions in both realms are likely
582 required, with the collaboration of terrestrial and aquatic ecologists and lighting professionals
583 (as occurred for the aforementioned turtle example), to successfully avoid terrestrial-aquatic
584 impacts.

585

586

587 **MITIGATING IMPACTS OF A BRIGHT FUTURE**

588 There is increasing recognition that conservation and management strategies should be
589 designed to account for cross-realm connections (e.g. Threlfall et al. 2021, Tulloch et al.
590 2021). A recent study developed a national-scale conservation framework that incorporated
591 linkages among the marine, freshwater and terrestrial realms, to select protected areas for
592 minimising the threats of both land-use and climate change (Tulloch et al. 2021). The cross-
593 realm approach resulted in changes to both terrestrial and marine priorities compared to when
594 connections among realms were not considered. The authors also argued that a cross-realm
595 approach allowed the identification of potential trade-offs and opportunity costs of
596 conservation versus ecological benefits, as well as the implementation of interventions with
597 multiple objectives (such as habitat management and biodiversity protection) (Tulloch et al.
598 2021).

599 Increasing the uptake of a cross-realm management approach requires increased and
600 improved communication between researchers, lighting practitioners, managers and
601 regulators that work within and across different realms. The creation of professional networks
602 is a great way to begin such conversations. In Australia, the Network for Ecological Research
603 on Artificial Light (NERAL; www.neralaus.com) was established to provide a platform to
604 connect researchers and practitioners working towards mitigating the impacts of light
605 pollution within and across realms. NERAL is a professional network of academic scientists
606 and consultants, with a wide range of expertise, including terrestrial and marine ecologists
607 and physiologists, and managers from local and federal government agencies. A primary aim
608 of the network is to increase communication between scientists and managers working on
609 different species, habitats and/or realms. This will allow: 1) managers to easily access
610 information crucial to developing and implementing interventions to prevent or mitigate light
611 pollution impacts, and 2) researchers to identify management priorities and provide evidence-

612 based information to shape management interventions. Networks that have a strong multi-
613 realm focus such as NERAL are important, as they enable a more holistic understanding of
614 issues related to ALAN. They can also provide an opportunity to develop standardised
615 methods for measuring light so that the impacts can be compared across realms. This holistic
616 approach can then be translated into the ongoing implementation of strategies to reduce
617 impacts of ALAN across terrestrial, marine and freshwater realms.

618

619 **BOX 1) LIGHT AS AN ECOLOGICAL TRAP**

620 Ecological traps form when animals are attracted into poor-quality habitats where
621 their fitness is compromised (Hale and Swearer 2016). ALAN can cause ecological traps by
622 influencing both the habitat selection decisions of animals and their fitness consequences.
623 The orb-web spiders and aquatic insect community case study presented here clearly
624 illustrates this – the adult stages of aquatic insects are attracted to artificial light where they
625 suffer higher mortality because of the high density of webs. This case study provides further
626 evidence of how ecological traps caused by ALAN can impact on cross-realm linkages. In
627 this case, ALAN strengthens the magnitude of cross-realm predator-prey interactions.
628 Specifically, the higher attraction and mortality of aquatic insects leads to increased aquatic-
629 to-terrestrial subsidy flux (e.g. Manfrin et al. 2017).

630 Artificial light can also interfere with the migratory behaviour of species that occupy
631 different realms as part of their life cycle. A well-known example of this is the impact of
632 ALAN on the dispersal behaviour of sea-turtle hatchlings. Nocturnally emerging hatchlings
633 are attracted to artificial lighting from coastal development. Crawling towards an artificial
634 light source can result in predation (Erb and Wyneken 2019), impair their ability to swim
635 offshore (Lorne and Salmon 2007), leading to reduced rates of offshore migration and rates
636 of transition between life stages (Wilson et al. 2019).

637 Lastly, ALAN could increase cross-realm rates of disease transmission due to its
638 impact on vector biology, such as biting mosquitoes. For example, in a recent study by Fyie
639 et al. (2021), artificial light masked natural daylength change which is the trigger for
640 diapause, meaning mosquitos remained reproductively active for longer and produced more
641 aquatic larvae. ALAN exposed mosquitos also had increased rates of blood feeding compared
642 to control mosquitos. Given the preference for humans to associate with artificially lit

643 environments at night, this suggests both changes in human and vector behaviour have
644 resulted in a largely unrecognized ecological trap for humans.

645

646 **BOX 2) CROSS-REALM EXPLOITATION OF RESOURCES USING ARTIFICIAL**
647 **LIGHT AT NIGHT**

648 Artificial light at night is known to attract and/or aggregate many organisms. This
649 effect can be exploited by predator species within and across realms, if, for example, a
650 terrestrial predator is exploiting an aggregation of aquatic organisms to a light source. One of
651 the best cross-realm examples of how ALAN can be used to exploit resources is the use of
652 artificial light by humans during night-time fishing.

653 The attraction of many fish and aquatic invertebrates to light has been known for
654 thousands of years, and artificial light has been used by humans to improve fishing efficacy
655 for centuries (Yami 1976). Light at night is known to attract small fish, insects and/or
656 plankton, which in turn attract larger predatory fishes and invertebrates (Becker et al. 2013),
657 or directly attracting target species through positive phototaxis, disorientation, or curiosity
658 (Marchesan et al. 2005). Historically, humans exploited this behaviour by lighting a fire on a
659 beach to attract fish into the shallows to facilitate harvest (e.g. by spearing or netting) (Yami
660 1976). Today, incandescent, fluorescent, metal halide, and LED above-water and underwater
661 lights are used for artisanal and industrialized fishing practices worldwide to increase harvest
662 (Solomon and Ahmed 2016, Nguyen and Winger 2019). In fact, certain fisheries cannot
663 operate effectively without the use of lights, such as the squid jigging fishery. Jigging for
664 squid dates back to antiquity in many parts of the world, however in the recent century, the
665 addition of artificial light to jigging gear has substantially increased landings due to the effect
666 of light at night on attracting and concentrating squid (Solomon and Ahmed 2016).

667 The effects of ALAN on fish attraction/aggregation are not lost on recreational
668 fishers; recreational fishers often target artificially lit areas for night fishing, as they know
669 certain target game species will follow baitfish into the illuminated areas (Cooke et al. 2017).
670 Urbanization has led to an increase in artificial light installations in coastal areas,
671 illuminating a substantial portion of shallow aquatic habitats at night (Davies et al. 2014,
672 Davies et al. 2016), and has therefore created ample opportunities for recreational fishers to
673 exploit artificial lighting (i.e. light pollution) to increase catch rates.

674 The increased harvest resulting from fishing practices using ALAN can lead to
675 overfishing and increased rates of bycatch in a fishery which may can have negative impacts
676 on fished populations (e.g. reduction in size and altered life-history traits) (Solomon and
677 Ahmed 2016) and thus ecological consequences for the marine or freshwater realms (e.g.
678 through trophic cascades). However, since responses to ALAN are species-specific, ALAN
679 can be used by humans to both increase fishing harvest and reduce catch rates of different
680 species. The use of artificial light has been recognized as a potential tool for bycatch
681 reduction in commercial fisheries, and therefore ALAN can also be exploited to mitigate
682 cross-realm impacts through minimizing effects of fishing on non-target organisms. Research
683 on the use of artificial light to reduce bycatch has demonstrated varying levels of success
684 (e.g. Hannah et al. 2015, Larsen et al. 2018, Lomeli et al. 2018) and is dependent on species
685 of interest, light properties tested, and proper placement/location of (often LED) lights within
686 the fishing gear. However, the use of artificial light to deter adult sea turtles has also proved
687 to be effective (e.g. Wang et al. 2010, Virgili et al. 2018) resulting in LED lights now widely
688 applied worldwide in pelagic gillnet fisheries to reduce sea turtle bycatch (Nguyen and
689 Winger 2019). This positive use of artificial light demonstrates that with species-specific
690 knowledge, it is possible to harness the effects of ALAN for positive impacts across realms.

691

692 **BOX 3: DISCREPANCIES IN LIGHT MEASUREMENTS**

693 A complicating factor influencing the ability of scientists to confidently predict the
694 impact of light on a sensitive receptor is the lack of an agreed upon standard method for
695 modelling, measuring and monitoring light or skyglow (e.g. Jechow and Hölker 2019,
696 Jechow et al. 2019, Kalinkat et al. 2021). Instrument types and applications vary widely:
697 instruments include luxmeters, spectrometers, and cameras which measure light emitted
698 directly from a source or light reflected from a surface, from overhead looking down on the
699 earth (satellite based) or from the ground looking up or horizontally across the landscape.
700 Limitations include: restrictions in the wavelengths they measure (i.e. they do not measure all
701 wavelengths across the entire visible spectrum), detection limits that are not low enough to
702 measure sky glow or intensities that elicit a biological response, highly technical instruments
703 requiring specialised knowledge to operate and maintain, and a wide range of different
704 measurement units.

705 Arguably, many of the existent ‘disparities’ are due to the fact that different
706 instruments are designed to measure different things, depending on the objectives of the
707 users. For example, studies aiming to measure large-scale environmental effects due to sky
708 glow will (and should) measure different variables (and consequently use different
709 instruments) than studies which the primary aim is to evaluate the effects of street-light on a
710 particular species of insect. Nevertheless, whenever possible, studies with similar objectives
711 and/or operating at similar spatial scales, should try to standardise measurements. Crucially it
712 is important to understand the operating limits of even the simplest instruments, as
713 instruments can be misused or used for an inappropriate environment (Longcore et al. 2020).
714 Similarly, the literature acknowledges that there are no conclusive intensity thresholds below
715 which artificial light is not harmful to species and habitats (Schroer et al. 2020), and even the
716 low intensity light characteristic of skyglow can affect organisms (Grubisic et al. 2019,

717 Kupprat et al. 2020).

718 Attempts to compare or standardise measurements across realms adds further
719 complications. For instance, while remote sensing techniques are commonly used as a best
720 proxy to quantify the amount of artificial light at night on terrestrial systems, there are serious
721 challenges associated with the use of this technology in water bodies/underwater (see the
722 extensive discussion in Jechow and Hölker 2019). Furthermore, different disciplines often
723 use different physical quantities and units for measuring light, creating confusion even among
724 experts (Jechow and Hölker 2019). For instance, much of the existing data on the quantity
725 and quality of light reaching both terrestrial and aquatic systems assess different physical
726 parameters (spectral irradiance, illuminance); have used several different instruments to
727 acquire measurements (e.g. SQM, luxmeter, spectrometer, digital camera); and, report
728 outcomes using different measurement units (lux, candela, magnitudes, Watts). Therefore, as
729 stated by Jechow and Hölker (2019), ‘there is no clear coherence between these
730 measurements, although each of them was well designed and conducted’. Cross-realm
731 assessment and management of light pollution is impeded by the discrepancies in
732 measurements of light pollution across systems and disciplines. However, standardization of
733 measurements across species level responses, systems, and realms of interest is incredibly
734 challenging, as measurements currently generally differ for valid, practical reasons, such as
735 the ecological and spatial scale of interest. This challenge highlights the value of cross-realm
736 and cross-discipline networks for developing solutions that allow efficient conservation and
737 management actions across species, habitats and realms.

738

739

740

741

742 **Acknowledgements**

743 This work stemmed from a workshop that was funded by a Flinders University Impact Seed
744 Funding awarded to E. Fobert and supported by the Sydney Institute of Marine Science
745 (SIMS). K. Robert and A. Dimovski are supported by a Net Zero Fund grant, sponsored by
746 Sonepar.

747

748 **Authorship contribution:**

749 All authors conceived the idea of the manuscript during a workshop hosted by SIMS. MMP
750 led the writing of the manuscript. MMP, EKF, TMJ, SES, and KAR wrote specific sections
751 of the manuscript. All authors contributed critically to the drafts and gave final approval for
752 publication.

753

754 **Data availability:** No data has been collected in preparation from this manuscript. All data
755 used is available in the literature.

756

757 **Conflict of interest:** The authors declare no conflict of interest.

758

759 **Ethics approval:** No ethics approval is needed for this manuscript.

760

761 **Consent for publication:** The authors have seen the manuscript and agree to its submission
762 for publication.

763

764

765 **REFERENCES**

766

767 Alvarez-Romero, J. G., V. M. Adams, R. L. Pressey, M. Douglas, A. P. Dale, A. A. Auge, D.
768 Ball, J. Childs, M. Digby, and R. Dobbs. 2015. Integrated cross-realm planning: A
769 decision-makers' perspective. *Biological Conservation* **191**:799-808.

770 Aulsebrook, A. E., F. Connelly, R. D. Johnsson, T. M. Jones, R. A. Mulder, M. L. Hall, A. L.
771 Vyssotski, and J. A. Lesku. 2020. White and Amber Light at Night Disrupt Sleep
772 Physiology in Birds. *Current Biology* **30**:3657-3663. e3655.

773 Baxter, C. V., K. D. Fausch, and W. Carl Saunders. 2005. Tangled webs: reciprocal flows of
774 invertebrate prey link streams and riparian zones. *Freshwater Biology* **50**:201-220.

775 Becker, A., A. K. Whitfield, P. D. Cowley, J. Järnegren, and T. F. Næsje. 2013. Potential
776 effects of artificial light associated with anthropogenic infrastructure on the
777 abundance and foraging behaviour of estuary-associated fishes. *Journal of Applied*
778 *Ecology* **50**:43-50.

779 Beger, M., H. S. Grantham, R. L. Pressey, K. A. Wilson, E. L. Peterson, D. Dorfman, P. J.
780 Mumby, R. Lourival, D. R. Brumbaugh, and H. P. Possingham. 2010. Conservation
781 planning for connectivity across marine, freshwater, and terrestrial realms. *Biological*
782 *Conservation* **143**:565-575.

783 Bennie, J., T. W. Davies, D. Cruse, and K. J. Gaston. 2016. Ecological effects of artificial
784 light at night on wild plants. *Journal of Ecology* **104**:611-620.

785 Bennie, J., T. W. Davies, D. Cruse, R. Inger, and K. J. Gaston. 2015. Cascading effects of
786 artificial light at night: resource-mediated control of herbivores in a grassland
787 ecosystem. *Philosophical Transactions of the Royal Society B: Biological Sciences*
788 **370**:20140131.

789 Bolton, D., M. Mayer-Pinto, G. Clark, K. Dafforn, W. Brassil, A. Becker, and E. Johnston.
790 2017. Coastal urban lighting has ecological consequences for multiple trophic levels
791 under the sea. *Science of the Total Environment* **576**:1-9.

792 Boughman, J. W. 2001. Divergent sexual selection enhances reproductive isolation in
793 sticklebacks. *Nature* **411**:944-948.

794 Bowyer, R. T., D. K. Person, and B. M. Pierce. 2005. Detecting top-down versus bottom-up
795 regulation of ungulates by large carnivores: implications for conservation of
796 biodiversity. *Large carnivores and the conservation of biodiversity*. Island Press,
797 Covelo:342-361.

798 Brännäs, E. 1995. First access to territorial space and exposure to strong predation pressure: a
799 conflict in early emerging Atlantic salmon (*Salmo salar* L.) fry. *Evolutionary Ecology*
800 **9**:411-420.

801 Brüning, A., F. Hölker, and C. Wolter. 2011. Artificial light at night: implications for early
802 life stages development in four temperate freshwater fish species. *Aquatic Sciences*
803 **73**:143-152.

804 Bugnot, A. B., G. C. Hose, C. J. Walsh, O. Floerl, K. French, K. A. Dafforn, J. Hanford, E.
805 C. Lowe, and A. K. Hahs. 2019. Urban impacts across realms: making the case for
806 inter-realm monitoring and management. *Science of the Total Environment* **648**:711-
807 719.

808 Campbell, S. 2007. *Ecological Specialisation and Conservation Biology of the Large-footed*
809 *Myotis: Myotis Macropus*. University of Melbourne, Department of Zoology.

810 Campbell, S. 2009. So long as it's near water: variable roosting behaviour of the large-footed
811 myotis (*Myotis macropus*). *Australian Journal of Zoology* **57**:89-98.

812 Campbell, S. 2011. Ecological specialisation and conservation of Australia's large-footed
813 myotis: a review of trawling bat behaviour. *The Biology and Conservation of*
814 *Australasian Bats*. (Eds B. Law, P. Eby, D. Lunney, and LF Lumsden.) pp:72-85.

815 Cinzano, P., F. Falchi, and C. D. Elvidge. 2001. The first world atlas of the artificial night
816 sky brightness. *Monthly Notices of the Royal Astronomical Society* **328**:689-707.

817 Collier, K. J., S. Bury, and M. Gibbs. 2002. A stable isotope study of linkages between
818 stream and terrestrial food webs through spider predation. *Freshwater Biology*
819 **47**:1651-1659.

820 Cooke, S. J., R. J. Lennox, S. D. Bower, A. Z. Horodysky, M. K. Treml, E. Stoddard, L. A.
821 Donaldson, and A. J. Danylchuk. 2017. Fishing in the dark: the science and
822 management of recreational fisheries at night. *Bulletin of Marine Science* **93**:519-538.

823 Davies, T. W., J. Bennie, and K. J. Gaston. 2012. Street lighting changes the composition of
824 invertebrate communities. *Biology Letters* **8**:764-767.

825 Davies, T. W., M. Coleman, K. M. Griffith, and S. R. Jenkins. 2015. Night-time lighting
826 alters the composition of marine epifaunal communities. *Biology Letters*
827 **11**:20150080.

828 Davies, T. W., J. P. Duffy, J. Bennie, and K. J. Gaston. 2014. The nature, extent, and
829 ecological implications of marine light pollution. *Frontiers in Ecology and the*
830 *Environment* **12**:347-355.

831 Davies, T. W., J. P. Duffy, J. Bennie, and K. J. Gaston. 2016. Stemming the tide of light
832 pollution encroaching into marine protected areas. *Conservation Letters* **9**:164-171.

833 DAWE. 2020. National Light Pollution Guidelines for Wildlife Including Marine Turtles,
834 Seabirds and Migratory Shorebirds, Commonwealth of Australia 2020. Australia
835 Government, Australia.

836 Dominoni, D., M. Quetting, and J. Partecke. 2013. Artificial light at night advances avian
837 reproductive physiology. *Proceedings of the Royal Society B: Biological Sciences*
838 **280**:20123017.

839 Dominoni, D., J. A. Smit, M. E. Visser, and W. Halfwerk. 2020. Multisensory pollution:
840 Artificial light at night and anthropogenic noise have interactive effects on activity
841 patterns of great tits (*Parus major*). *Environmental Pollution* **256**:113314.

842 Durrant, J., L. M. Botha, M. P. Green, and T. M. Jones. 2018. Artificial light at night
843 prolongs juvenile development time in the black field cricket, *Teleogryllus*
844 *commodus*. *Journal of Experimental Zoology Part B: Molecular and Developmental*
845 *Evolution* **330**:225-233.

846 Dwyer, P. 1970. Foraging behaviour of the Australian large-footed *Myotis* (Chiroptera).
847 Edensor, T. 2017. From light to dark: Daylight, illumination, and gloom. U of Minnesota
848 Press.

849 Eisenbeis, G., C. Rich, and T. Longcore. 2006. Artificial night lighting and insects: attraction
850 of insects to streetlamps in a rural setting in Germany. *Ecological consequences of*
851 *artificial night lighting* **2**:191-198.

852 Endler, J. A. 1993. The color of light in forests and its implications. *Ecological Monographs*
853 **63**:1-27.

854 Erb, V., and J. Wyneken. 2019. Nest-to-surf mortality of loggerhead sea turtle (*Caretta*
855 *caretta*) hatchlings on Florida's east coast. *Frontiers in Marine Science* **6**:271.

856 Falchi, F., P. Cinzano, D. Duriscoe, C. C. Kyba, C. D. Elvidge, K. Baugh, B. A. Portnov, N.
857 A. Rybnikova, and R. Furgoni. 2016. The new world atlas of artificial night sky
858 brightness. *Science advances* **2**:e1600377.

859 Field, R. D., and L. Parrott. 2017. Multi-ecosystem services networks: A new perspective for
860 assessing landscape connectivity and resilience. *Ecological Complexity* **32**:31-41.

861 Fobert, E. K., K. Burke da Silva, and S. E. Swearer. 2019. Artificial light at night causes
862 reproductive failure in clownfish. *Biology Letters* **15**:20190272.

863 Fyie, L. R., M. M. Gardiner, and M. E. Meuti. 2021. Artificial light at night alters the
864 seasonal responses of biting mosquitoes. *Journal of Insect Physiology* **129**:104194.

- 865 Gallaway, T., R. N. Olsen, and D. M. Mitchell. 2010. The economics of global light
866 pollution. *Ecological Economics* **69**:658-665.
- 867 Gaston, K. J., T. W. Davies, J. Bennie, and J. Hopkins. 2012. Reducing the ecological
868 consequences of night-time light pollution: options and developments. *Journal of*
869 *Applied Ecology* **49**:1256-1266.
- 870 Gaston, K. J., T. W. Davies, S. L. Nedelec, and L. A. Holt. 2017. Impacts of artificial light at
871 night on biological timings. *Annual Review of Ecology, Evolution, and Systematics*
872 **48**:49-68.
- 873 Gaston, K. J., J. P. Duffy, S. Gaston, J. Bennie, and T. W. Davies. 2014. Human alteration of
874 natural light cycles: causes and ecological consequences. *Oecologia* **176**:917-931.
- 875 Gende, S. M., R. T. Edwards, M. F. Willson, and M. S. Wipfli. 2002. Pacific salmon in
876 aquatic and terrestrial ecosystems: Pacific salmon subsidize freshwater and terrestrial
877 ecosystems through several pathways, which generates unique management and
878 conservation issues but also provides valuable research opportunities. *Bioscience*
879 **52**:917-928.
- 880 Giakoumi, S., V. Hermoso, S. B. Carvalho, V. Markantonatou, M. Dagys, T. Iwamura, W. N.
881 Probst, R. J. Smith, K. L. Yates, and V. Almpnidou. 2019. Conserving European
882 biodiversity across realms. *Conservation Letters* **12**:e12586.
- 883 Gorecki, V., M. Rhodes, and S. Parsons. 2020. Roost selection in concrete culverts by the
884 large-footed myotis (*Myotis macropus*) is limited by the availability of microhabitat.
885 *Australian Journal of Zoology*.
- 886 Grubisic, M., A. Haim, P. Bhusal, D. M. Dominoni, K. Gabriel, A. Jechow, F. Kupprat, A.
887 Lerner, P. Marchant, and W. Riley. 2019. Light pollution, circadian photoreception,
888 and melatonin in vertebrates. *Sustainability* **11**:6400.
- 889 Grubisic, M., G. Singer, M. C. Bruno, R. H. van Grunsven, A. Manfrin, M. T. Monaghan,
890 and F. Hölker. 2017. Artificial light at night decreases biomass and alters community
891 composition of benthic primary producers in a sub-alpine stream. *Limnology and*
892 *Oceanography* **62**:2799-2810.
- 893 Haddock, J. K. 2019. Effects of artificial lighting on insectivorous bat communities in urban
894 ecosystems.
- 895 Hale, R., J. J. Piggott, and S. E. Swearer. 2017. Describing and understanding behavioral
896 responses to multiple stressors and multiple stimuli. *Ecology and Evolution* **7**:38-47.
- 897 Hale, R., and S. E. Swearer. 2016. Ecological traps: current evidence and future directions.
898 *Proceedings of the Royal Society B: Biological Sciences* **283**:20152647.
- 899 Hannah, R. W., M. J. Lomeli, and S. A. Jones. 2015. Tests of artificial light for bycatch
900 reduction in an ocean shrimp (*Pandalus jordani*) trawl: strong but opposite effects at
901 the footrope and near the bycatch reduction device. *Fisheries research* **170**:60-67.
- 902 Harewood, A., and J. Horrocks. 2008. Impacts of coastal development on hawksbill hatchling
903 survival and swimming success during the initial offshore migration. *Biological*
904 *Conservation* **141**:394-401.
- 905 Haywood, J. C., W. J. Fuller, B. J. Godley, J. D. Shutler, S. Widdicombe, and A. C.
906 Broderick. 2019. Global review and inventory: how stable isotopes are helping us
907 understand ecology and inform conservation of marine turtles. *Marine Ecology*
908 *Progress Series* **613**:217-245.
- 909 Hernández, A. L. M., and B. I. van Tussenbroek. 2014. Patch dynamics and species shifts in
910 seagrass communities under moderate and high grazing pressure by green sea turtles.
911 *Marine Ecology Progress Series* **517**:143-157.
- 912 Hilderbrand, G. V., T. A. Hanley, C. T. Robbins, and C. C. Schwartz. 1999. Role of brown
913 bears (*Ursus arctos*) in the flow of marine nitrogen into a terrestrial ecosystem.
914 *Oecologia* **121**:546-550.

915 Hölker, F., C. Wurzbacher, C. Weißenborn, M. T. Monaghan, S. I. Holzhauer, and K.
916 Premke. 2015. Microbial diversity and community respiration in freshwater sediments
917 influenced by artificial light at night. *Philosophical Transactions of the Royal Society*
918 *B: Biological Sciences* **370**:20140130.

919 Hopkins, G. R., K. J. Gaston, M. E. Visser, M. A. Elgar, and T. M. Jones. 2018. Artificial
920 light at night as a driver of evolution across urban–rural landscapes. *Frontiers in*
921 *Ecology and the Environment* **16**:472-479.

922 Horváth, G., G. Kriska, P. Malik, and B. Robertson. 2009. Polarized light pollution: a new
923 kind of ecological photopollution. *Frontiers in Ecology and the Environment* **7**:317-
924 325.

925 Horváth, G., A. Lerner, and N. Shashar. 2014. Polarized light and polarization vision in
926 animal sciences. Springer.

927 Jechow, A., and F. Hölker. 2019. How dark is a river? Artificial light at night in aquatic
928 systems and the need for comprehensive night-time light measurements. *Wiley*
929 *Interdisciplinary Reviews: Water* **6**:e1388.

930 Jechow, A., C. Kyba, and F. Hölker. 2019. Beyond all-sky: assessing ecological light
931 pollution using multi-spectral full-sphere fisheye lens imaging. *Journal of Imaging*
932 **5**:46.

933 Jones, M., A. Laurila, N. Peuhkuri, J. Piironen, and T. Seppä. 2003. Timing an ontogenetic
934 niche shift: responses of emerging salmon alevins to chemical cues from predators
935 and competitors. *Oikos* **102**:155-163.

936 Juday, C., W. H. Rich, G. I. Kemmerer, and A. Mann. 1932. Limnological studies of Karluk
937 lake, Alaska, 1926-1930. US Government Printing Office.

938 Kalinkat, G., M. Grubisic, A. Jechow, R. H. van Grunsven, S. Schroer, and F. Hölker. 2021.
939 Assessing long-term effects of artificial light at night on insects: what is missing and
940 how to get there. *Insect Conservation and Diversity* **14**:260-270.

941 Kriska, G., B. Bernath, R. Farkas, and G. Horvath. 2009. Degrees of polarization of reflected
942 light eliciting polarotaxis in dragonflies (Odonata), mayflies (Ephemeroptera) and
943 tabanid flies (Tabanidae). *Journal of Insect Physiology* **55**:1167-1173.

944 Kühne, J. L., R. H. van Grunsven, A. Jechow, and F. Hölker. 2021. Impact of Different
945 Wavelengths of Artificial Light at Night on Phototaxis in Aquatic Insects. *Integrative*
946 *and Comparative Biology* **61**:1182-1190.

947 Kummu, M., H. De Moel, P. J. Ward, and O. Varis. 2011. How close do we live to water? A
948 global analysis of population distance to freshwater bodies. *PloS one* **6**:e20578.

949 Kupprat, F., F. Hölker, and W. Kloas. 2020. Can skyglow reduce nocturnal melatonin
950 concentrations in Eurasian perch? *Environmental Pollution* **262**:114324.

951 Kyba, C. C., T. Kuester, A. S. De Miguel, K. Baugh, A. Jechow, F. Hölker, J. Bennie, C. D.
952 Elvidge, K. J. Gaston, and L. Guanter. 2017. Artificially lit surface of Earth at night
953 increasing in radiance and extent. *Science advances* **3**:e1701528.

954 Land, M. F., and D.-E. Nilsson. 2012. Animal eyes. Oxford University Press.

955 Larsen, R. B., B. Herrmann, M. Sistiaga, J. Brčić, J. Brinkhof, and I. Tatone. 2018. Could
956 green artificial light reduce bycatch during Barents Sea Deep-water shrimp trawling?
957 *Fisheries research* **204**:441-447.

958 Law, B., and C. Urquhart. 2000. Diet of the Large-footed Myotis *Myotis Mcropus* at A Forest
959 Stream Roost in Northern New South Wales. *Australian mammalogy* **22**:121-124.

960 Levin, L. A., D. F. Boesch, A. Covich, C. Dahm, C. Erséus, K. C. Ewel, R. T. Kneib, A.
961 Moldenke, M. A. Palmer, and P. Snelgrove. 2001. The function of marine critical
962 transition zones and the importance of sediment biodiversity. *Ecosystems* **4**:430-451.

963 Lomeli, M. J., S. D. Groth, M. T. Blume, B. Herrmann, and W. W. Wakefield. 2018. Effects
964 on the bycatch of eulachon and juvenile groundfish by altering the level of artificial

illumination along an ocean shrimp trawl fishing line. *ICES Journal of Marine Science* **75**:2224-2234.

967 Longcore, T., D. Duriscoe, M. Aubé, A. Jechow, C. Kyba, and K. L. Pendoley. 2020.
968 Commentary: Brightness of the night sky affects loggerhead (*Caretta caretta*) sea
969 turtle hatchling misorientation but not nest site selection. *Frontiers in Marine Science*
970 **7**.

971 Lorne, J. K., and M. Salmon. 2007. Effects of exposure to artificial lighting on orientation of
972 hatchling sea turtles on the beach and in the ocean. *Endangered Species Research*
973 **3**:23-30.

974 Manfrin, A., D. Lehmann, R. H. van Grunsven, S. Larsen, J. Syväranta, G. Wharton, C. C.
975 Voigt, M. T. Monaghan, and F. Hölker. 2018. Dietary changes in predators and
976 scavengers in a nocturnally illuminated riparian ecosystem. *Oikos* **127**:960-969.

977 Manfrin, A., G. Singer, S. Larsen, N. Weiß, R. H. van Grunsven, N.-S. Weiß, S. Wohlfahrt,
978 M. T. Monaghan, and F. Hölker. 2017. Artificial light at night affects organism flux
979 across ecosystem boundaries and drives community structure in the recipient
980 ecosystem. *Frontiers in Environmental Science* **5**:61.

981 Marchesan, M., M. Spoto, L. Verginella, and E. A. Ferrero. 2005. Behavioural effects of
982 artificial light on fish species of commercial interest. *Fisheries research* **73**:171-185.

983 Mayer-Pinto, M., K. A. Dafforn, and E. L. Johnston. 2019. A decision framework for coastal
984 infrastructure to optimise biotic resistance and resilience in a changing climate.
985 *Bioscience* **69**:833-843.

986 McLay, L., M. Green, and T. Jones. 2017. Chronic exposure to dim artificial light at night
987 decreases fecundity and adult survival in *Drosophila melanogaster*. *Journal of Insect*
988 *Physiology* **100**:15-20.

989 McNaughton, E. J., K. J. Gaston, J. R. Beggs, D. N. Jones, and M. C. Stanley. 2021. Areas of
990 ecological importance are exposed to risk from urban sky glow: Auckland, Aotearoa-
991 New Zealand as a case study. *Urban Ecosystems*:1-12.

992 Mehner, T. 2012. Diel vertical migration of freshwater fishes—proximate triggers, ultimate
993 causes and research perspectives. *Freshwater Biology* **57**:1342-1359.

994 Meyer, S., G. Wegener, K. G. Lloyd, A. Teske, A. Boetius, and A. Ramette. 2013. Microbial
995 habitat connectivity across spatial scales and hydrothermal temperature gradients at
996 Guaymas Basin. *Frontiers in microbiology* **4**.

997 Miller, C. R., B. T. Barton, L. Zhu, V. C. Radeloff, K. M. Oliver, J. P. Harmon, and A. R.
998 Ives. 2017. Combined effects of night warming and light pollution on predator–prey
999 interactions. *Proceedings of the Royal Society B: Biological Sciences* **284**:20171195.

1000 Naiman, R. J., and H. Decamps. 1997. The ecology of interfaces: riparian zones. *Annual*
1001 *Review of Ecology and Systematics* **28**:621-658.

1002 Nguyen, K. Q., and P. D. Winger. 2019. Artificial light in commercial industrialized fishing
1003 applications: a review. *Reviews in Fisheries Science & Aquaculture* **27**:106-126.

1004 Oppedal, F., T. Dempster, and L. H. Stien. 2011. Environmental drivers of Atlantic salmon
1005 behaviour in sea-cages: a review. *Aquaculture* **311**:1-18.

1006 Parkinson, E., J. Lawson, and S. D. Tiegs. 2020. Artificial light at night at the terrestrial-
1007 aquatic interface: Effects on predators and fluxes of insect prey. *PloS one*
1008 **15**:e0240138.

1009 Pauley, S. M. 2004. Lighting for the human circadian clock: recent research indicates that
1010 lighting has become a public health issue. *Medical hypotheses* **63**:588-596.

1011 Pérez Vega, C., K. M. Zielinska-Dabkowska, and F. Hölker. 2021. Urban Lighting Research
1012 Transdisciplinary Framework—A Collaborative Process with Lighting Professionals.
1013 *International Journal of Environmental Research and Public Health* **18**:624.

- 1014 Perkin, E. K., F. Hölker, J. S. Richardson, J. P. Sadler, C. Wolter, and K. Tockner. 2011. The
1015 influence of artificial light on stream and riparian ecosystems: questions, challenges,
1016 and perspectives. *Ecosphere* **2**:1-16.
- 1017 Perkin, E. K., F. Hölker, and K. Tockner. 2014. The effects of artificial lighting on adult
1018 aquatic and terrestrial insects. *Freshwater Biology* **59**:368-377.
- 1019 Pittman, J., and D. Armitage. 2016. Governance across the land-sea interface: a systematic
1020 review. *Environmental Science & Policy* **64**:9-17.
- 1021 Reimchen, T. E. 2000. Some ecological and evolutionary aspects of bear-salmon interactions
1022 in coastal British Columbia. *Canadian Journal of Zoology* **78**:448-457.
- 1023 Riley, W., B. Bendall, M. Ives, N. Edmonds, and D. Maxwell. 2012. Street lighting disrupts
1024 the diel migratory pattern of wild Atlantic salmon, *Salmo salar* L., smolts leaving
1025 their natal stream. *Aquaculture* **330**:74-81.
- 1026 Riley, W., P. Davison, D. Maxwell, and B. Bendall. 2013. Street lighting delays and disrupts
1027 the dispersal of Atlantic salmon (*Salmo salar*) fry. *Biological Conservation* **158**:140-
1028 146.
- 1029 Riley, W., P. Davison, D. Maxwell, R. Newman, and M. Ives. 2015. A laboratory experiment
1030 to determine the dispersal response of Atlantic salmon (*Salmo salar*) fry to street light
1031 intensity. *Freshwater Biology* **60**:1016-1028.
- 1032 Robert, K. A., J. A. Lesku, J. Partecke, and B. Chambers. 2015. Artificial light at night
1033 desynchronizes strictly seasonal reproduction in a wild mammal. *Proceedings of the*
1034 *Royal Society B: Biological Sciences* **282**:20151745.
- 1035 Robertson, B., G. Kriska, V. Horvath, and G. Horvath. 2010. Glass buildings as bird feeders:
1036 urban birds exploit insects trapped by polarized light pollution. *Acta Zoologica*
1037 *Academiae Scientiarum Hungaricae* **56**:283-293.
- 1038 Rodríguez, A., G. Burgan, P. Dann, R. Jessop, J. J. Negro, and A. Chiaradia. 2014. Fatal
1039 attraction of short-tailed shearwaters to artificial lights. *PloS one* **9**:e110114.
- 1040 Rodríguez, A., P. Dann, and A. Chiaradia. 2017. Reducing light-induced mortality of
1041 seabirds: high pressure sodium lights decrease the fatal attraction of shearwaters.
1042 *Journal for Nature Conservation* **39**:68-72.
- 1043 Sanders, D., E. Frago, R. Kehoe, C. Patterson, and K. J. Gaston. 2020. A meta-analysis of
1044 biological impacts of artificial light at night. *Nature ecology & evolution*:1-8.
- 1045 Schlaepfer, M. A., M. C. Runge, and P. W. Sherman. 2002. Ecological and evolutionary
1046 traps. *Trends in Ecology & Evolution* **17**:474-480.
- 1047 Schroer, S., B. J. Huggins, C. Azam, and F. Hölker. 2020. Working with inadequate tools:
1048 legislative shortcomings in protection against ecological effects of artificial light at
1049 night. *Sustainability* **12**:2551.
- 1050 Schulte-Römer, N., J. Meier, E. Dannemann, and M. Söding. 2019. Lighting Professionals
1051 versus Light Pollution Experts? Investigating Views on an Emerging Environmental
1052 Concern. *Sustainability* **11**:1696.
- 1053 Senzaki, M., J. R. Barber, J. N. Phillips, N. H. Carter, C. B. Cooper, M. A. Ditmer, K. M.
1054 Frstrup, C. J. McClure, D. J. Mennitt, and L. P. Tyrrell. 2020. Sensory pollutants
1055 alter bird phenology and fitness across a continent. *Nature* **587**:605-609.
- 1056 Shimoda, M., and K.-i. Honda. 2013. Insect reactions to light and its applications to pest
1057 management. *Applied Entomology and Zoology* **48**:413-421.
- 1058 Solomon, O. O., and O. O. Ahmed. 2016. Fishing with light: Ecological consequences for
1059 coastal habitats. *International Journal of Fisheries and Aquatic Studies* **4**:474-483.
- 1060 Straka, T. M., P. E. Lentini, L. F. Lumsden, B. A. Wintle, and R. van der Ree. 2016. Urban
1061 bat communities are affected by wetland size, quality, and pollution levels. *Ecology*
1062 *and Evolution* **6**:4761-4774.

- 1063 Sullivan, S. M. P., K. Hossler, and L. A. Meyer. 2019. Artificial lighting at night alters
1064 aquatic-riparian invertebrate food webs. *Ecological Applications* **29**:e01821.
- 1065 Suthers, R. A. 1965. Acoustic orientation by fish-catching bats. *Journal of Experimental*
1066 *Zoology* **158**:319-347.
- 1067 Swearer, S. E., R. L. Morris, L. T. Barrett, M. Sievers, T. Dempster, and R. Hale. 2021. An
1068 overview of ecological traps in marine ecosystems. *Frontiers in Ecology and the*
1069 *Environment*.
- 1070 Szaz, D., G. Horvath, A. Barta, B. A. Robertson, A. Farkas, A. Egri, N. Tarjanyi, G. Racz,
1071 and G. Kriska. 2015. Lamp-lit bridges as dual light-traps for the night-swarming
1072 mayfly, *Ephoron virgo*: interaction of polarized and unpolarized light pollution. *PloS*
1073 *one* **10**:e0121194.
- 1074 Taylor, P. D., L. Fahrig, K. Henein, and G. Merriam. 1993. Connectivity is a vital element of
1075 landscape structure. *Oikos* **68**:571-573.
- 1076 Thompson, D., and M. Fenton. 1982. Echolocation and feeding behaviour of *Myotis adversus*
1077 (*Chiroptera: Vespertilionidae*). *Australian Journal of Zoology* **30**:543-546.
- 1078 Threlfall, C. G., E. M. Marzinelli, A. Ossola, A. B. Bugnot, M. J. Bishop, L. Lowe, S.
1079 Imberger, S. Myers, P. D. Steinberg, and K. A. Dafforn. 2021. Towards cross-realm
1080 management of urban ecosystems. *Frontiers in Ecology and the Environment*.
- 1081 Thums, M., S. D. Whiting, J. Reisser, K. L. Pendoley, C. B. Pattiaratchi, M. Proietti, Y.
1082 Hetzel, R. Fisher, and M. G. Meekan. 2016. Artificial light on water attracts turtle
1083 hatchlings during their near shore transit. *Royal Society open science* **3**:160142.
- 1084 Tulloch, V. J., S. Atkinson, H. P. Possingham, N. Peterson, S. Linke, J. R. Allan, A. Kaiye,
1085 M. Keako, J. Sabi, and B. Suruman. 2021. Minimizing cross-realm threats from land-
1086 use change: A national-scale conservation framework connecting land, freshwater and
1087 marine systems. *Biological Conservation* **254**:108954.
- 1088 Villalobos-Jimenez, G., A. Dunn, and C. Hassall. 2016. Dragonflies and damselflies
1089 (*Odonata*) in urban ecosystems: a review. *European Journal of Entomology* **113**:217-
1090 232.
- 1091 Virgili, M., C. Vasapollo, and A. Lucchetti. 2018. Can ultraviolet illumination reduce sea
1092 turtle bycatch in Mediterranean set net fisheries? *Fisheries research* **199**:1-7.
- 1093 Wang, J. H., S. Fislser, and Y. Swimmer. 2010. Developing visual deterrents to reduce sea
1094 turtle bycatch in gill net fisheries. *Marine Ecology Progress Series* **408**:241-250.
- 1095 Webb, T. J. 2012. Marine and terrestrial ecology: unifying concepts, revealing differences.
1096 *Trends in Ecology & Evolution* **27**:535-541.
- 1097 Williams, C., and B. Singh. 1951. Effect of moonlight on insect activity. *Nature* **167**:853-853.
- 1098 Willmott, N. J., J. Henneken, M. A. Elgar, and T. M. Jones. 2019. Guiding lights: Foraging
1099 responses of juvenile nocturnal orb-web spiders to the presence of artificial light at
1100 night. *Ethology* **125**:289-297.
- 1101 Willmott, N. J., J. Henneken, C. J. Selleck, and T. M. Jones. 2018. Artificial light at night
1102 alters life history in a nocturnal orb-web spider. *PeerJ* **6**:e5599.
- 1103 Wilson, P., M. Thums, C. Pattiaratchi, M. Meekan, K. Pendoley, R. Fisher, and S. Whiting.
1104 2018. Artificial light disrupts the nearshore dispersal of neonate flatback turtles
1105 *Natator depressus*. *Marine Ecology Progress Series* **600**:179-192.
- 1106 Wilson, P., M. Thums, C. Pattiaratchi, S. Whiting, K. Pendoley, L. C. Ferreira, and M.
1107 Meekan. 2019. High predation of marine turtle hatchlings near a coastal jetty.
1108 *Biological Conservation* **236**:571-579.
- 1109 Witherington, B. E., and K. A. Bjorndal. 1991. Influences of artificial lighting on the seaward
1110 orientation of hatchling loggerhead turtles *Caretta caretta*. *Biological Conservation*
1111 **55**:139-149.
- 1112 Yami, B. 1976. Fishing with light. *FAO fishing manuals*.

1113 Zapata, M. J., S. M. P. Sullivan, and S. M. Gray. 2019. Artificial Lighting at Night in
1114 Estuaries—Implications from Individuals to Ecosystems. *Estuaries and coasts* **42**:309-
1115 330.
1116